1	Modelling methane emissions and grain yields for a double-rice system in									
2	Southern China with DAYCENT and DNDC models									
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14										
15	Abstract									
16	Methane (CH <sub>4</sub> ) is an important greenhouse gas (GHG) that contributes to climate									
17	change and one of its major sources is rice cultivation. The main aim of this paper was									
18	to compare two well-established biogeochemical models, namely Daily Century									
19	(DAYCENT) and DeNitrification-DeComposition (DNDC) for estimating CH4									
20	emissions and grain yields for a double-rice cropping system with tillage practice									

19 (DAYCENT) and DeNitrification-DeComposition (DNDC) for estimating CH<sub>4</sub> 20 emissions and grain yields for a double-rice cropping system with tillage practice 21 and/or stubble incorporation in the winter fallow season in Southern China. Both 22 models were calibrated and validated using field measured data from November 2008 23 to November 2014. The calibrated models performed effectively in estimating the 24 daily CH<sub>4</sub> emission pattern (correlation coefficient, r = 58-63, p < 0.001), but model 25 efficiency (EF) values were higher in stubble incorporation treatments, with and

without winter tillage (treatments S and WS) (EF = 0.22-0.28) than that in winter 26 tillage without stubble incorporation treatment (W) (EF = -0.06-0.08). We 27 recommend that algorithms for the impacts of tillage practice on CH<sub>4</sub> emission should 28 be improved for both models. DAYCENT and DNDC also estimated rice yields for all 29 treatments without a significant bias. Our results showed that tillage practice in the 30 winter fallow season (treatments WS and W) significantly decreased annual CH<sub>4</sub> 31 emissions, by 13-37% (p < 0.05) for measured values, 15-20% (p < 0.05) for 32 DAYCENT-simulated values, and 12-32% (p < 0.05) for DNDC-simulated values, 33 34 respectively, compared to no-till practice (treatments S), but had no significant impact on grain yields. 35

36 Keywords: Methane; Management practices; DAYCENT model; DNDC model;
37 Double-rice cropping system; China

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#### 39 **1. Introduction**

Methane (CH<sub>4</sub>) is a powerful greenhouse gas (GHG), the atmosphere amount of 40 which has more than doubled since pre-industrial times, and approximately 60% of 41 CH<sub>4</sub> originates from anthropogenic sources (Nisbet et al., 2019; UNEP and CCAC, 42 2021). Rice paddy fields are a major source of CH<sub>4</sub> emissions, which are responsible 43 for 8–11% (5–38 Tg CH<sub>4</sub> yr<sup>-1</sup>) of global anthropogenic CH<sub>4</sub> emissions (Shukla et al., 44 45 2019; Saunois et al., 2020). Quantification of CH<sub>4</sub> emissions from rice paddy soils is necessary for developing mitigation options and policies. However, accurate 46 estimation of CH<sub>4</sub> emissions is a great challenge due to the time consuming and 47 expensive field flux measurements. Consequently, process-based models for 48 estimating CH<sub>4</sub> emissions have been developed to complement physical experiments 49 by employing computers to calculate the likely outcomes of different physical 50

51 phenomena (Giltrap et al., 2010).

Simulation models allow complex interactions and real-world problems to be 52 examined in a cost- and time-effective way (Giltrap et al., 2010; Cheng et al., 2013). 53 DAYCENT and DeNitrification-DeComposition (DNDC) are two popular ecological 54 process-based models to simulate methane (CH<sub>4</sub>) emissions from rice paddy fields in 55 China (Li et al., 2006; Cheng et al., 2014; Zhang et al., 2019; Wang et al., 2021). 56 57 Cheng et al. (2013) developed and evaluated the DAYCENT CH<sub>4</sub> module using total 97 rice paddy sites across China, with an overall r of 0.83 for model predictions vs. 58 59 measurements. In addition, the DNDC model has been corroborated by many CH<sub>4</sub> emission datasets from Chinese rice fields, and the simulated values are generally in 60 good agreement with the observed CH<sub>4</sub> field emissions (Zhang et al., 2002; Li et al., 61 62 2006; Zhang et al., 2019; Zhao et al., 2020; Wang et al., 2021).

China is the largest rice producer in the world and is also trying to increase rice 63 grain yield by improving rice cultivation management, while at the same time, 64 minimizing CH<sub>4</sub> emissions from rice paddy. Rice is a one of the primary cereal crops 65 in China, with an area of about 30 million ha (FAO, 2020). Double rice is the common 66 cropping system in China, accounting for >40% of the total harvested area and 67 emitting about 50% of the total CH<sub>4</sub> emission from rice paddy fields in China (Zhang 68 et al., 2011; Chen et al., 2013). The double rice cropping system typically consists of 69 70 one winter fallow season, and two rice growing seasons each year. Considerable research has been conducted on improving field management in the double-rice 71 cropping system to mitigate CH<sub>4</sub> emissions while maintaining optimal rice yields. 72 73 These have mainly focused on the fertilization rate and method (Tang et al., 2020; Fu et al., 2021; Wang H. et al., 2021), irrigation management method (Li et al., 2020; 74 Zeng and Li, 2020), and tillage management (Chen et al., 2021; Wang X. et al., 2021). 75

Tillage after rice harvest in the winter fallow season can play key role in CH<sub>4</sub> 76 emissions. It is beneficial for rainwater to run through into the subsoil, and thus 77 reduce rainwater accumulation in the winter fallow season. Consequently, it would 78 directly reduce CH<sub>4</sub> emission during off-rice season because of a less anaerobic 79 environment in the topsoil (Zhang et al., 2016). Moreover, it could also indirectly 80 inhibit CH<sub>4</sub> emissions during the following rice growth seasons. For example, tillage 81 82 incorporates rice residues into the soil during winter fallow season, and soil microorganisms accelerate the decomposition of organic matter and thereby facilitate 83 CH<sub>4</sub> production and emissions (Pandey et al., 2012; Hussain et al., 2015). 84 Subsequently, it would reduce the carbon substrate for methanogenesis during the 85 following rice seasons, and thus decrease CH<sub>4</sub> emissions (Yang et al., 2018). 86

87 As a representative region of the double-rice cropping system, Jiangxi Province has the largest rice area; about 11% of total rice area in China (Yearbook, 2014) and 88 emits substantial quantities of CH4. However, to the best of our knowledge, the 89 process-based DNDC and DAYCENT models have not previously been calibrated and 90 evaluated for a double-rice cropping system with different management in the winter 91 fallow season in China. Moreover, little research has been done to compare different 92 process-based models in simulating CH<sub>4</sub> emissions. Therefore, the main objective of 93 this study was to compare the results from two well-established biogeochemical 94 models (DAYCENT and DNDC) for estimating CH4 emissions and crop yields from a 95 double rice system in Jiangxi Province, southeast China from November 2008 to 96 November 2014 under three different managements in the winter fallow season. This 97 study will improve process understanding and enhance further applicability of 98 DAYCENT and DNDC models for predicting CH<sub>4</sub> emissions from the Chinese paddy 99 rice ecosystem. 100

#### 102 2. Materials and Methods

#### 103 2.1. Experimental site and treatments

This field experiment was conducted at Yingtan City, Jiangxi Province, China 104 (28°15'N, 116°55'E) for 6 years from November 2008 to November 2014. This region 105 is a typical double-rice cropping cultivation area, with one winter fallow season and 106 107 two rice growing seasons each year. The selected soil is classified as a typical Haplaquept (18.2% clay, 31.3% silt, 50.5% sand), with its initial properties as follow: 108 SOC 16.2 g kg<sup>-1</sup>, soil total nitrogen 1.43 g kg<sup>-1</sup>, bulk density 1.12 g cm<sup>-3</sup>, pH (H<sub>2</sub>O) 109 4.74. The detailed site description and soil parameters were reported by Yang et al. 110 (2018). The daily air temperature (°C) and precipitation (mm) were collected from 111 weather station at the study site (Fig. S1). The average annual temperature and total 112 precipitation were 18.2°C and 194.2 cm, respectively. The monthly mean air 113 temperature and rainfall from 2008 and 2014 at the field site are presented in Table 114 S1. 115

Compared with rice stubble incorporation during the rice season, applying rice 116 stubble during the fallow season produces much lower CH<sub>4</sub> emissions (Yan et al., 117 2009). Additionally, soil tillage with rice stubble incorporation in the winter fallow 118 season has been reported to reduce CH<sub>4</sub> emission relative to rice stubble incorporation 119 120 just before rice transplanting (Zhang et al., 2016; Yang et al., 2018). In this study, three treatments were laid out in the winter fallow season with three replicates in a 121 fully randomized block design: rice stubble incorporation without winter tillage (S), 122 winter tillage with rice stubble incorporation (WS), and winter tillage without rice 123 stubble incorporation (W). Fresh rice stubble was left standing in the fields after late 124 rice harvest in treatments S and WS, with a dry weight of 2.5-4.0 t ha<sup>-1</sup> (about 30 cm 125

long), while stubble was moved out of field after late rice harvest in treatment W. Noextra straw/stubble was incorporated in the following rice seasons.

Generally, ploughing is the traditional tillage practice in the local area, with tillage occurring before the transplantation of early- and late-rice. For better cultivation, all experimental plots (S, WS and W) were ploughed before the transplantation of early- and late-rice rice without any rice stubble/straw incorporation. The winter tillage plots (treatments WS and W) were ploughed again as soon as the late rice had been harvested. The tillage operation (up to 20 cm soil depth) was the same for all tillage practices.

Local rice cultivars, Zhongzao 33 and Nongxiang 98, were planted in the 135 following early-rice and late-rice seasons, respectively. Seeds were sown in the 136 seeding nursery and then transplanted to the experimental plots at the third and fourth 137 leaf stage. The early rice seedlings were transplanted in middle or late April and 138 harvested in middle or late July, and then late rice seedlings were transplanted 139 immediately after the early rice harvest and harvested in November or December from 140 2009 and 2014 (Table 1). For each rice season, the total amount of nitrogen (N) and 141 potassium (K) fertilizers applied were 180 kg N ha<sup>-1</sup> and 150 kg K ha<sup>-1</sup>, respectively. 142 These fertilizers were applied at three different times as basal, tillering and panicle 143 initiation fertilizer with a ratio of 5:3:2 and 3:4:3, respectively. Phosphorus (P) 144 fertilizer was applied as a basal fertilizer at a rate of 75 kg P ha<sup>-1</sup>. 145

For water management, flooding was initiated 2-4 days before early-rice transplanting, drained after tillering fertilization application for 5-8 days midseason aeration, re-flooded for two or three weeks, then subjected to drying-wetting alternation (with a cycle of 5-day drying and 5 day-wetting) until roughly 1-2 weeks of a dry period before early rice harvest. During the late-rice season, the water management was similar to that during the early-rice season but the duration of the dry period before late rice harvest was roughly 3-5 weeks. A detailed schedule of the field management, including soil tillage, rice cultivation and water management, is presented in Table 1.

#### 155 2.2. Field measurements and GHG emissions

The CH<sub>4</sub> fluxes were measured using a static chamber (Ma et al., 2009), every 2 to 6 days over the rice seasons, and every 7 to 10 days over the winter fallow seasons in 15 min intervals. The yield of early- and late-rice grain was determined at harvest in each plot by subtracting a moisture content of 0.14 g H<sub>2</sub>O g<sup>-1</sup> fresh weight. The details of measurement information for daily CH<sub>4</sub> flux and yield were described by Yang et al. (2018).

162 GHG emissions (kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>) based on CH<sub>4</sub> emission were estimated 163 using global warming potential (GWP) (CO<sub>2</sub>-eq) for CH<sub>4</sub> over a 100-year time span 164 (Eq (1)). The GWP for CH<sub>4</sub> is 28 over a 100-year time span (IPCC, 2021):

166 To determine the emission intensity of production, GHG emission per unit of 167 crop yield was calculated (Eq (2)):

168 yield-scaled 
$$GHG = GHG/$$
 (early rice yield + late rice yield) (2)

169

#### 170 2.3. Model descriptions and simulations

We used two process-based ecosystem models, DAYCENT and DNDC, developed to simulate soil carbon and nitrogen dynamic in plant-soil system (Parton et al., 1998; Li, 2000; Gilhespy et al., 2014). Model concept and mechanisms are described in greater detail elsewhere for DAYCENT (Del Grosso et al., 2001; Cheng et al., 2013; Begum et al., 2019), and DNDC (Li et al., 1994; Li et al., 2006). Daily weather data, plant, soil and management data including N fertilizer, watermanagement and tillage are needed as inputs for both models.

With an understanding of the processes of CH<sub>4</sub> production, oxidation and 178 emission, a methanogenesis sub-model for the DAYCENT model was developed for 179 predicting methane fluxes dynamics in rice paddy soils by Cheng et al. (2013). 180 Rice-DAYCENT simulates plant production, soil organic matter (SOM) 181 182 decomposition, soil hydrology and thermal regimes. The methanogenesis sub-model simulates CH<sub>4</sub> emissions based on methanogenic substrate derived from SOM 183 184 decomposition and root rhizodeposition, and associated influences of redox potential (Eh) and soil temperature (Huang et al., 1998; Cheng et al., 2013). As described in 185 Cheng et al. (2013), the decomposition of organic matter in soil was simulated by 186 DAYCENT model through heterotrophic respiration using three kinetically defined 187 active, slow and passive pools. The amount of carbon added to the soil through 188 rhizodeposition was simulated using a simplified linear equation with root carbon 189 production estimated in the plant production sub-model. The influence of Eh was 190 simulated under flooding and drainage, respectively. Only part of the CH<sub>4</sub> produced in 191 the process of methanogenesis is emitted to atmosphere because about 40-90% of CH<sub>4</sub> 192 is oxidized to CO<sub>2</sub> by methanotrophs at aerobic-anaerobic interfaces (Huang et al., 193 1998; Chen et al., 2013). The pathway of CH<sub>4</sub> from the paddy soil into the atmosphere 194 195 occurs in various ways: via aerenchyma in the plant (90%), via ebullition (10%) or via diffusion through the soil and water layer (1%) (Groot et al., 2003). The 196 methanogenesis sub-model adopted the approach proposed by Huang et al. (2008, 197 2004) to simulate CH<sub>4</sub> emissions through the rice plant and ebullition. The simulation 198 of CH<sub>4</sub> emission rates through the rice plant was based on the CH<sub>4</sub> production rate, 199 and the fraction of CH<sub>4</sub> emitted via rice. The algorithm for simulating CH<sub>4</sub> emissions 200

through ebullition was based on CH<sub>4</sub> production rate, soil temperature, and root
biomass. The CH<sub>4</sub> oxidation model was based on field capacity, bulk density, soil
temperature, water-filled pore space and volumetric soil water content.

The DNDC model was modified by adding a series of anaerobic process for 204 simulating the carbon cycle and CH<sub>4</sub> emission in rice paddy field as described in Li et 205 al. (2000; 2004). The DNDC model accommodates two components. The first 206 207 component consists of three main sub-models as follow: the soil climate sub-model calculating soil temperature, moisture and Eh profiles; the plant growth sub-model 208 209 simulating crop biomass accumulation and partitioning; the decomposition sub-model simulating concentration of substrates, i.e., dissolved organic carbon and NH<sub>4</sub><sup>+</sup>, 210 nitrogen oxides. The second component, namely the fermentation sub-model, predicts 211 212 the CH<sub>4</sub> fluxes dynamics from plant-soil systems. For example, CH<sub>4</sub> production rate was simulated using kinetical equations based on available carbon concentration and 213 temperature as soon as the simulated Eh reaches -150 mV or lower. In addition, CH<sub>4</sub> 214 oxidation rate was simulated using a function of soil CH<sub>4</sub> concentration and Eh. 215 DNDC models simulated CH<sub>4</sub> emissions through plant aerenchyma and ebullition, 216 respectively, based on CH<sub>4</sub> concentration, soil temperature and soil porosity. 217

218

#### 219 2.4. Model calibrations and sensitivity analyses

This study investigated the suitability of the DAYCENT and DNDC models for estimating CH<sub>4</sub>, crop yield for typical double rice paddy field in Southern China. This double rice cropping system in our study consists of a 4- or 5-month long winter fallow season, followed by early rice (grown from April to July), and then late rice planted immediately after the early rice harvest (grown from July to November/December). The DAYCENT model was calibrated on crop yield / annual

CH<sub>4</sub> emissions for the site using the measured data from the control treatment S. 226 Model calibration for crop yield / annual CH<sub>4</sub> emissions was done by optimizing the 227 crop parameters of radiation use efficiency (PRDX), optimum temperature (PPDF(1)), 228 and the fraction of  $CO_2$  from soil respiration used to produce  $CH_4$  (Table 2), as 229 suggested by previous studies (Cheng et al., 2014; Begum et al., 2019). The parameter 230 values were modified until the DAYCENT model matched measured grain 231 232 yield/annual CH<sub>4</sub> emission values from the control treatment S. The calibrated model was then used to run those for another two treatments WS and W from November 233 234 2008 to November 2014.

Similarly, the DNDC model was also calibrated on crop yield/annual CH4 235 emissions for the site using the measured data from the control with treatment S. 236 237 Model calibration for crop yields and CH<sub>4</sub> emissions was done by optimizing a combination of different crop growth parameters, including maximum biomass 238 production, biomass fraction, biomass C/N ratio, thermal degree days (Table 3), as 239 suggested by Zhang et al. (2019) and Abdalla et al. (2020). Crop parameter input 240 default values were tested until the DNDC model matched the measured grain 241 yield/annual CH<sub>4</sub> emission values from the control treatment S. The calibrated model 242 was then used to run those for another two treatments WS and W from November 243 2008 to November 2014. 244

The sensitivity of DAYCENT and DNDC and the attribution of CH<sub>4</sub> and early-/late-rice grain yields to different input parameters were investigated to quantify the effects of these parameters on the CH<sub>4</sub> emissions and grain yields (Smith and Smith, 2007; Cheng et al., 2013; Wang et al., 2021). The baseline scenario was composed based on the treatment S. Only one parameter was changed at a time and all the other kept constant. Simulations were run to assess how CH<sub>4</sub> and grain yields were affected by average daily temperature (increased/decreased by a range from  $-2 \,^{\circ}C$  to +2  $\,^{\circ}C$ ), initial SOC content (decreased/increased by a rang from -50% to +50%), soil pH (decreased/increased by a range from -1 to +1) and the amounts of N fertilizer (decreased/increased by a rang from -50% to +50%).

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#### 256 2.5. Statistical methods

The models were validated by comparing measured and simulated values. Based on the statistical routines provide in MODEVAL (Smith et al., 1997; Smith & Smith, 2007), the total difference between measured and simulated values was assessed by calculating the root mean square error (RMSE, Eq. (3)), relative RMSE (rRMSE, Eq (4)), relative deviation (RD, Eq(5)):

263 
$$rRMSE = \frac{RMSE}{\overline{M}} \times 100$$
 (4)

264 
$$RD = \frac{M_i - S_i}{M_i} \times 100$$
(5)

Where *Si* is the simulated value, *Mi* is the measured value, n is the number of measured values, and  $\overline{M}$  is the average of the measured values. The rRMSE can compare between different models whose errors are measured in the different units, and a low rRMSE often indicates a strong predictive power.

The DAYCENT and DNDC models' accuracies were evaluated by calculating modelling efficiency (EF, Eq (6)). EF provides a comparison of the efficiency of the chosen model compared to describing the data as the mean of the measurements (Yang et al., 2014):

273 
$$EF = 1 - \frac{\sum_{i=1}^{n} (Si - Mi)^{2}}{\sum_{i=1}^{n} (Mi - \overline{M})^{2}}$$
(6)

Values of EF can be positive or negative values. Specifically, a positive value shows that the simulated values describe the trend in the measured data better than the mean of the measurements, and closer to 1 suggests a better modelling efficiency. A negative value indicates that the simulated values describe the data less well than a mean of the measurements.

The sample correlation coefficient (r) was used (Eq. (7)) to test for association between the simulated and measured values (Smith et al., 1997).

281 
$$\mathbf{r} = \frac{\sum_{i=1}^{n} (M_i - \overline{M})(S_i - \overline{S})}{\sqrt{\left[\sum_{i=1}^{n} (M_i - \overline{M})^2\right]} \sqrt{\left[\sum_{i=1}^{n} (S_i - \overline{S})^2\right]}}$$
(7)

All the statistical analyses were conducted in R version 3.4.0 (Team, 2008) and Minitab (Minitab, Limited Liability Company, USA), and a Map was created using Original (Origin Lab Corporation, USA).

285

286 **3. Results** 

#### 287 3.1. Models calibration and sensitivity analyses

The DAYCENT and DNDC models were calibrated by adjusting the combination of crop parameters as shown in Tables 2 and 3 to enhance their performances in simulating CH<sub>4</sub> emissions and grain yields. The calibrated DAYCENT and DNDC model accurately simulated the measured annual CH<sub>4</sub> emissions, early and late rice yields for the control with treatment S from November 2008 to November 2014 (Table 4).

The sensitivity of the DAYCENT- and DNDC-models to the essential input parameters (i.e. SOC content, soil pH, the N fertilizer rate and air temperature) for simulating annual CH<sub>4</sub> emission and grain yield of double-cropping rice system was
tested. As shown in Fig. 1, DAYCENT was more sensitive to changes in SOC content
and soil pH than the other parameters, whilst the DNDC was more sensitive to
changes in air temperature and N fertilizer. For grain yields, neither model was
sensitive to change in air temperature, but DNDC was very sensitive to changes in N
fertilizer rate and SOC content (Fig. 1).

302

# 303 3.2. Performance of DAYCENT and DNDC models in simulating CH<sub>4</sub> emissions 304 and rice grain yields

#### 305 *3.2.1. CH*<sup>4</sup> *emissions*

Fig. 2 shows that, for all treatments, DAYCENT- and DNDC-simulated daily 306 CH<sub>4</sub> emissions pattern were generally consistent with the measured CH<sub>4</sub> flux 307 dynamics. The daily CH<sub>4</sub> emissions for all three treatments increased under 308 continuous flooding, with the highest peak measured at about 3-5 weeks after the 309 early-rice transplanting and 2–4 weeks after late-rice transplanting. Thereafter, daily 310 CH<sub>4</sub> emissions dramatically decreased after midseason aeration. An emission peak 311 occurred again after re-flooding, particularly in the early-rice season. CH<sub>4</sub> emissions 312 always showed a lower peak in the treatment W, observed both in simulations and 313 measurements. As shown in Table 5, DAYCENT and DNDC models performed better 314 when simulating treatments S and WS, with a lower rRMSE (i.e., 124-129) and 315 higher EF values (i.e., 0.22–0.28), than treatment W (i.e., 140-150 and -0.07–0.08, 316 respectively), but all three treatments showed significant correlations of simulated 317 versus measured daily emission values (r = 58-63, p < 0.001). 318

The annual CH<sub>4</sub> emissions simulated by DAYCENT and DNDC models were also generally similar to the measured annual values for all three treatments (Table 4).

The measured average annual CH<sub>4</sub> emissions were 175, 152, and 111 kg C ha<sup>-1</sup> for the 321 treatment S, WS and W, respectively (Table 4). Correspondingly, the DAYCENT- and 322 DNDC-simulated average annual CH<sub>4</sub> emissions were 173, 148 and 138 kg C ha<sup>-1</sup>, 323 and 173, 153 and 117 kg C ha<sup>-1</sup>, respectively. Both the observed and simulated results 324 showed significantly lower (p < 0.05) annual CH<sub>4</sub> emissions from the treatment W 325 than from the treatment S. Over the six annual rotation cycles from November 2008 to 326 327 November 2014, the measured annual CH<sub>4</sub> emission was not significantly different within years for treatment S, while significantly decreased from the first rotation year 328 329 of 2008-2009 to final rotation year of 2013-2014 for treatments WS and W (Fig.3).

As shown in Fig.4, winter tillage (treatments WS and W) decreased the seasonal 330 CH<sub>4</sub> emission for early rice season from -36 to -15% for measured values (p<0.05), 331 from -26 to -17% for DAYCENT-simulated values, and from -38 to -13% for 332 DNDC-simulated values. Similarly, the seasonal CH<sub>4</sub> emissions for late rice season 333 also decreased from -40 to -14% for measured values, from -18 to -14% for 334 DAYCENT-simulated values, and from -28 to -11% for DNDC-simulated values. 335 By contrast, the tillage in winter fallow season (treatments WS and W) increased the 336 seasonal CH<sub>4</sub> emission by 31-87% for measured values (p<0.05) and 9-36% for 337 DAYCENT-simulated (p < 0.05) compared to no-till treatment (treatment S). 338

339 *3.2.2. Rice yields* 

The DAYCENT and DNDC models estimated grain yield for all treatments effectively (Table 4). As shown in Fig. 5, the correlation coefficient (r) of simulated against measured yields of both early and late rice season were 0.90, 0.85 and 0.92 by DAYCENT model (p < 0.001), which were higher than the values of 0.82 (p < 0.01), 0.67 (p < 0.05) and 0.58 by the DNDC model, for treatments S, WS and S, respectively.

On average, the measured yields were 6.3, 6.6, and 6.5 t  $ha^{-1}$  over early rice, and 346 6.4, 6.5, and 6.3 t  $ha^{-1}$  over late rice, for the treatments S. WS and W. respectively 347 (Table 4). Correspondingly, the DAYCENT-simulated average yields were 6.4, 6.4, 348 and 6.4 t  $ha^{-1}$  over early rice, and 6.3, 6.3, and 6.3 t  $ha^{-1}$  over late rice; 349 DNDC-simulated average yields were 6.1, 6.7, and 7.0 t  $ha^{-1}$  over early rice, and 6.6, 350 6.8, and 6.7 t  $ha^{-1}$  over late rice, respectively. Overall, the grain yields were not 351 significantly different among the three treatments, observed both in measurements and 352 simulations (Table 4). 353

Over the six annual rotation cycles from November 2008 to November 2014, the annual yields were not significantly different within most years, except in the rotation year of 2009-2010. The lower annual yield in 2010 was due to the flood damage, resulting in the delaying of late rice transplanting, thus reducing the rice grain yields (Fig.3).

359

## 360 3.2.3. Yield-scaled GHG emissions

Compared with the treatment S, measured yield-scaled GHG emissions were lower by 17% for treatment WS and by 38% for treatment W (p < 0.01) (Table 3). Similarly, simulated yield-scaled GHG emissions were lower by 15% and 16% with treatment WS, by 21% and 37% with treatment W (p < 0.01) for DAYCENT and DNDC, respectively.

366

## 367 **4. Discussion**

#### 368 4.1. Model calibration and sensitivity analysis

369 In this study, calibration and validation of DAYCENT and DNDC models was 370 required because of differences in the Chinese rice cultivars and climates (Cheng et al., 2013; Wang et al., 2021). However, the adopted parameters for calibration between
DAYCENT and DNDC models are different due to differences in the crop growth and
CH<sub>4</sub> algorithms in the two models (Li, 2000; Cheng et al., 2013).

Sensitivity analysis was also used to evaluate the response of the simulated 374 results to the variation in the input parameters. We utilized the calibrated DAYCENT 375 and DNDC models to test how CH<sub>4</sub> emission and rice grain yield were influenced by 376 377 soil properties, climate factors and N fertilizer application rates. As the CH<sub>4</sub> algorithm is implemented in different ways, the results indicate the robustness and uncertainty of 378 379 the different processes. While the models showed good performances on aerobic systems, impacts of management changes and mitigation strategies, the diverse 380 management on the considered sites will allow the models to be challenged on these 381 aspects as well. For both of CH<sub>4</sub> emission and grain yields, DAYCENT and DNDC 382 models were not sensitive to the same parameters as shown in Fig. 1, which may be 383 due to differences in the algorithms of the methanogenesis sub-model (Li, 2000; 384 Cheng et al., 2013), thus resulting in the differences of dominant factors influencing 385 CH<sub>4</sub> emissions, with the effects of other factors being overshadowed by the influence 386 of the dominant factors (Wang et al., 2021). 387

For simulating CH<sub>4</sub> emissions, the DAYCENT model is more sensitive to 388 changes in initial SOC content. The initial SOC content determined the amount of 389 390 carbon substrate for methanogenic bacteria, for CH<sub>4</sub> production and also emissions (Conrad, 2007). Therefore, annual CH<sub>4</sub> emissions changed with a change in the initial 391 SOC content in the same direction under otherwise identical conditions (Fig. 1). By 392 contrast, the DNDC model was less sensitive to the changes of initial SOC content 393 (Fig. 1), which was also reported by Wang et al. (2021). This can be explained by 394 differences in the calculation of available C from SOM decomposition between the 395

two process models. Moreover, DAYCENT and DNDC models have a different way of representing initial SOC. For example, the initial SOC stock  $(g m^{-2})$  at 20 cm soil depth was required to define the initial soil organic matter pools in DAYCENT model, but initial SOC content (kg kg<sup>-1</sup>) at 10 cm soil depth was required in DNDC model. Therefore, when the same changes of initial SOC content were applied, DAYCENT and DNDC models have different relative changes of initial carbon stock input, thus different changes of available C concentration.

Decreased soil pH (pH < 4.7, under acidic conditions) significantly decreased 403 404 annual CH<sub>4</sub> emissions in DAYCENT model, but increased pH slightly increased CH<sub>4</sub> emissions, which is related to the soil pH thresholds effecting decomposition rate in 405 the model. When pH value decreases especially from ~5 to 3, the decomposition rate 406 407 dramatically reduces in the DAYCENT model, thereby significantly decreasing CH4 emissions. By contrast, when soil pH value increases from 5 to 7, the decomposition 408 rate barely changes because, it is close to the maximum rate in the DAYCENT model. 409 Cheng et al. (2013) also showed that the performance in simulating in CH<sub>4</sub> emission 410 by DAYCENT was mainly controlled by the initial SOC content and soil pH. 411 However, for DNDC, the annual CH<sub>4</sub> emission was not sensitive to the changes of soil 412 pH, but very sensitive to air temperature (Wang et al., 2021). As shown in Li (2000), 413 the effect of temperature on CH<sub>4</sub> production rates in DNDC is based on an 414 exponential function, and when temperature increase, the temperature effect becomes 415 larger directly. Moreover, DNDC simulates CH<sub>4</sub> fluxes diffusion through ebullition to 416 atmosphere using a simplified linear equation with temperature. Therefore, this is 417 probably why a significantly effect of temperature on CH<sub>4</sub> emissions was observed in 418 DNDC model. By contrast, in DAYCENT model, the algorithm for calculating 419 transport CH<sub>4</sub> through ebullition was based on a natural logarithm function with 420

421 temperature, thus there is barely changes of temperature effects when temperature
422 increase/decrease within 2 °C (Cheng et al., 2013).

For simulating yields, the DAYCENT model was slightly sensitive to changes of 423 air temperature, which may be due to the saturation effect above 30 °C for rice paddy 424 in the model. In the test site, the average values of maximum temperature in rice 425 reproductive period were 30.29-34.58 °C during June to September, therefore the 426 427 simulated yields only slightly changed with air temperature changes. In contrast, the DNDC model was also slightly sensitive to changes of air temperature, but very 428 429 sensitive to the changes of N fertilizer rate and initial SOC content. In the plant growth sub model of DNDC, N uptake by crop is the key process linking crop growth, 430 and the availability of NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> in soil profile is one of main controlling factors 431 on N uptake rate (Li et al. 1994). Therefore, changes of N fertilizer rate directly affect 432 the concentration of  $NH_4^+$  and  $NO_3^-$  in the model, and thereby influence rice plant 433 growth and yields as well. On the other hand, calculating NH<sub>4</sub><sup>+</sup> concentration from N 434 fertilizer in DNDC model is also controlled by the concentration of soluble C from 435 decomposition sub model, which is why changes of initial SOC content in DNDC 436 directly affects the rice plant growth and grain yields (Li et al., 1992). 437

438

#### 439 *4.2. Evaluation of DAYCENT and DNDC models*

440 *4.2.1. CH<sub>4</sub> emissions* 

Simulation of substrate C available under different water and field management is crucial for predicting CH<sub>4</sub> emissions accurately by DAYCENT (Cheng et al., 2013) and DNDC (Li, 2007). Large CH<sub>4</sub> emissions were simulated at the middle growth stage in the month of May for early rice, and at the early growth stage in the month of July-August for late rice, when carbohydrates derived from plants was greater, and

soil Eh was lower due to continuous flooding conditions after rice transplantation in 446 this study. A clear CH<sub>4</sub> peak was also simulated during the re-flooding period after 447 midseason aeration in the month of May-June for early rice and August-September for 448 late rice. This could be due to re-flooding cutting off the oxygen supply from the air 449 into soil and decreasing soil Eh, thus benefiting methanogenetic activity (Cai et al., 450 2000). Correspondingly, both the DAYCENT and DNDC models simulated relatively 451 452 lower soil Eh during re-flooding period after midseason aeration, with on average values of -193, -192 and -190 and -185, -174 and -173 mV for treatment S, WS 453 454 and W, respectively.

A difference between seasonal simulated and measured CH<sub>4</sub> emissions was 455 observed in this study, especially in the winter fallow season. In the test sites, field 456 plots were fallow in the winter season with soil being undrained after late rice harvest, 457 which were often flooded after rain (Zhang et al., 2016), hence providing favourable 458 anaerobic conditions for CH<sub>4</sub> production. Compared with DNDC model, DAYCENT 459 accurately estimated the seasonal CH<sub>4</sub> emissions during the winter fallow season, 460 mainly due to better simulating the water condition during the winter fallow season. 461 However, the DNDC model runs without setting flooding conditions during the winter 462 fallow season because there is not a suitable corresponding flood setting option in the 463 model, thereby resulting in underestimated seasonal CH<sub>4</sub> emissions during winter 464 fallow season. But the seasonal CH<sub>4</sub> emissions during the winter fallow season 465 contributed, on average, around 2% to the annual CH<sub>4</sub> emissions observed in 466 measured and DAYCENT-simulated values, hence it had small effects on the 467 estimation of annual CH<sub>4</sub> emissions. On the other hand, DNDC underestimated the 468 seasonal CH<sub>4</sub> emissions from early rice seasons while slightly overestimating 469 emissions from late rice seasons for all treatment S, WS and W, which may be due to 470

the sensitivity of the DNDC model to air temperature changes. Slightly lower air
temperatures were found in the month of May-June (i.e., 22.9–26.0 °C) compared to
July-September (i.e., 25.6–29.6 °C) in this study, which also led to a CH<sub>4</sub> emission
peak for early and late rice season, respectively.

The response of CH<sub>4</sub> emissions to the incorporation of stubble was influenced 475 by the winter tillage. Winter tillage (treatments WS and W) significantly increased 476 477 CH<sub>4</sub> emission by 31-87% for measured values during the winter fallow season relative to no winter tillage (treatment S) (Fig. 4), in agreement with previous 478 479 measurements from a single-cropping rice field in northeast China (Liang et al., 2007). By contrast, it significantly decreased CH<sub>4</sub> emissions during the following early- and 480 late-rice seasons by -36 to -15% (Fig. 4), in agreement with our early field 481 observation (Yang et al., 2017), and previous measurements from a single-cropping 482 rice field in southern Brazil (Bayer et al., 2015). 483

The impact of winter tillage practices was satisfactorily replicated by both 484 DAYCENT and DNDC models. Compared to no-tillage in the winter fallow season, 485 winter tillage promotes the decomposition of rice stubble, which creates an anaerobic 486 soil environment suitable for methanogenic activity because of oxygen consumption, 487 and thereby enhanced observed/simulated CH4 emissions in the winter fallow season 488 (Zhang et al., 2015; Yang et al., 2018). By contrast, as the easily decomposable 489 portion of the rice stubble has largely been decomposed during the whole winter 490 fallow season, the positive effect of the remaining rice stubble (a less-decomposable 491 part of organic matter) on CH<sub>4</sub> production and emissions is greatly reduced during the 492 following seasons (Watanabe and Kimura, 1998; Bayer et al., 2015). 493

494

495 *4.2.2. Rice yields* 

An adequate simulation of yield is of key importance to accurately predict CH<sub>4</sub> 496 emissions for process-based models of plant-soil systems because carbohydrate 497 498 exudation from roots, the major labile carbon source driving CH<sub>4</sub> emissions, is closely related to rice plant biomass (Cheng et al., 2013). Both models simulating rice yields 499 performed effectively after calibration in this study. Significant positive correlations 500 of simulated against measured rice yields were observed in this study, with r values of 501 502 0.85–0.92 for DAYCENT, and 0.67–0.82 for DNDC (Fig. 5). Similar previous studies in China were also able to simulate rice yield adequately using the DAYCENT 503 504 (Stehfest et al., 2007; Cheng et al., 2013) and DNDC models (Zhang et al., 2019; Zhao et al., 2020). It is crucial the key growth processes (i.e. plant production and 505 allocation of net primary production, mineralization/immobilization, and nutrients 506 507 uptake by plant) are well represented in the approaches of the DAYCENT and DNDC models (Li et al., 1994; Cheng et al., 2013). 508

Tillage and/or stubble incorporation in winter fallow season did not impact rice 509 vields significantly (Table 4 and Fig. 3). In the DNDC and DAYCENT models, once 510 the soil is ploughed, decomposition rates of soil organic matter would be directly 511 increased due to the changes in soil structure and aeration conditions (Li et al., 1994; 512 Cheng et al., 2013). As for stubble incorporation after harvest, SOM would increase 513 by a certain percentage in DAYCENT and DNDC models. However, changes SOM 514 515 (i.e. soil C content) would not have a direct effect on simulation of yield, especially in DAYCENT. Moreover, only 15% of leaf and stem was assumed to be left in field after 516 harvest in the DNDC model, which might have less impact on total SOM, and thereby 517 rice yields. 518

519

#### 520 4.2.3. Yield-scaled GHG emissions

Compared with the treatment S, annual CH<sub>4</sub> emissions were clearly lower in the 521 treatments of WS and W, observed in both field measured and simulated results (Table 522 4). Similar measured results from a single-cropping rice field in northeast China were 523 reported by Liang et al. (2007). Additionally, maintaining rice paddy yield has always 524 been given priority before implementation of alternative management practice (Liu et 525 al., 2016). In this study, no significant differences in rice paddy yields were observed 526 527 among three treatments over the six years, consequently, annual yield-scaled GHG emissions were lower in the treatments of WS and W compared with treatment S for 528 529 both model simulated and field measured results (Table 4). Similar findings were shown by Zhang et al. (2016) and Yang et al. (2018). This indicates that the tillage 530 practice in the winter fallow season could be a potential strategy for reducing annual 531 GHG emissions without a significant impact on grain yield in double rice cropping 532 533 systems.

534

#### 535 **5. Conclusions**

This study has provided an insight into the differences of model performance 536 between DNDC and DAYCENT in simulating CH<sub>4</sub> emission from a double-rice 537 cropping system in Southern China. Both models were able to effectively estimate 538 daily CH<sub>4</sub> emission patterns and grain yields across all treatments from November 539 540 2008 to November 2014. Compared with the DNDC model, DAYCENT simulated the seasonal CH<sub>4</sub> emissions during winter fallow seasons better, mainly due to better 541 reflecting the water conditions in the real field for winter fallow seasons. Moreover, 542 the high sensitivity of the DNDC model to air temperature results in imperfectly 543 estimated seasonal CH<sub>4</sub> emissions for early and late rice seasons. As observed in the 544 simulations of both models and field measurements, the tillage practice in the winter 545

fallow season could be a potential strategy for reducing annual CH<sub>4</sub> emissions without a significantly impacting grain yield in double rice cropping systems. Further measurements of emissions for tillage and/or stubble incorporation in the winter fallow season are recommended before implementing the model outcomes.

550

#### 551 Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

554

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## 562 **References**

Abdalla, M., Song, X., Ju, X., Topp, C.F.E., Smith, P., 2020. Calibration and
validation of the DNDC model to estimate nitrous oxide emissions and crop
productivity for a summer maize-winter wheat double cropping system in Hebei,
China. Environ. Pollut. 262, 114199.

Bayer, C., Zschornack, T., Pedroso, G.M., da Rosa, C.M., Camargo, E.S., Boeni, M.,
Marcolin, E., dos Reis, C.E.S., dos Santos, D.C., 2015. A seven-year study on
the effects of fall soil tillage on yield-scaled greenhouse gas emission from flood
irrigated rice in a humid subtropical climate. Soil Tillage Res. 145, 118–125.

571	Begum, K., Kuhnert, M., Yeluripati, J.B., Ogle, S.M., Parton, W.J., Williams, S.A.,
572	Pan, G., Cheng, K., Ali, M.A., Smith, P., 2019. Modelling greenhouse gas
573	emissions and mitigation potentials in fertilized paddy rice fields in Bangladesh.
574	Geoderma 341, 206–215.

\*\*\*\*\*\*

- Cai, Z.C., Tsuruta, H., Minami, K., 2000. Methane emission from rice fields in China:
  measurements and influencing factors. J. Geophys. Res. Atmos. 105,
  17231–17242.
- Chen, H., Zhu, Q., Peng, C., Wu, N., Wang, Y., Fang, X., Jiang, H., Xiang, W.,
  Chang, J., Deng, X., Yu, G., 2013. Methane emissions from rice paddies natural
  wetlands, lakes in China: Synthesis new estimate. Glob. Chang. Biol. 19, 19–32.
- Chen, Z., Zhang, H., Xue, J., Liu, S., Chen, F., 2021. A nine-year study on the effects
  of tillage on net annual global warming potential in double rice-cropping systems
  in Southern China. Soil Tillage Res. 206, 104797.
- Cheng, K., Ogle, S.M., Parton, W.J., Pan, G., 2014. Simulating greenhouse gas
  mitigation potentials for Chinese Croplands using the DAYCENT ecosystem
  model. Glob. Chang. Biol. 20, 948–962.
- Cheng, K., Ogle, S.M., Parton, W.J., Pan, G., 2013. Predicting methanogenesis from
  rice paddies using the DAYCENT ecosystem model. Ecol. Modell. 261, 19–31.
- Conrad, R., 2007. Microbial ecology of methanogens and methanotrophs. Adv. Agron.
  96, 1–63.
- 591 Del Grosso, S.J., Parton, W.J., Mosier, A.R., Hartman, M.D., Brenner, J., Ojima, D.S.,
- Schimel, D.S., 2001. Simulated interaction of carbon dynamics and nitrogen
  trace gas fluxes using the DAYCENT model. Model. carbon nitrogen Dyn. soil
  Manag. 303–332.
- 595 FAO, 2019. FAOSTAT rice production data/China mainland (2019) [WWW

- 596 Document]. https:// faostat3.fao.org.
- Giltrap, D.L., Li, C., Saggar, S., 2010. DNDC: A process-based model of greenhouse
  gas fluxes from agricultural soils. Agric. Ecosyst. Environ. 136, 292–300.
- Huang, Y., Sass, R.L., Fisher, F.M., 1998. A semi-empirical model of methane
  emission from flooded rice paddy soils. Glob. Chang. Biol. 4, 247–268.
- Hussain, S., Peng, S., Fahad, S., Khaliq, A., Huang, J., Cui, K., Nie, L., 2015. Rice
- management interventions to mitigate greenhouse gas emissions: a review.
  Environ. Sci. Pollut. Res. 22, 3342–3360.
- 604 IPCC 2021. Climate Change 2021: The Physical Science Basis. Contribution of
- Working Group I to the Sixth Assessment Report of the Intergovernmental Panelon Climate Change. (Cambridge University Press).
- 607 Li, C., 2007. Quantifying greenhouse gas emissions from soils: Scientific basis and
- modeling approach. Soil Sci. Plant Nutr. 53, 344–352.
- Li, C., Frolking, S., Frolking, T.A., 1992. A model of nitrous oxide evolution from
  soil driven by rainfall events: 1. Model structure and sensitivity. J. Geophys. Res.
  Atmos. 97, 9759–9776.
- Li, C., Frolking, S., Harriss, R., 1994. Modeling carbon biogeochemistry in
  agricultural soils. Global Biogeochem. Cycles 8, 237–254.
- Li, C., Mosier, A., Wassmann, R., Cai, Z., Zheng, X., Huang, Y., Tsuruta, H.,
  Boonjawat, J., Lantin, R., 2004. Modeling greenhouse gas emissions from
  rice-based production systems: Sensitivity and upscaling. Global Biogeochem.
  Cycles 18, 1–19.
- Li, C., Salas, W., DeAngelo, B., Rose, S., 2006. Assessing Alternatives for Mitigating
  Net Greenhouse Gas Emissions and Increasing Yields from Rice Production in
  China Over the Next Twenty Years. J. Environ. Qual. 35, 1554–1565.

- Li, C.S., 2000. Modeling trace gas emissions from agricultural ecosystems. Methane
  Emiss. from major rice Ecosyst. asia 259–276.
- Li, L., Li, F., Dong, Y., 2020. Greenhouse Gas Emissions and Global Warming
  Potential in Double-Cropping Rice Fields as Influenced by Two Water-Saving
  Irrigation Modes in South China. J. Soil Sci. Plant Nutr. 20, 2617–2630.
- 626 Liu, S.-L., Pu, C., Ren, Y.-X., Zhao, X.-L., Zhao, X., Chen, F., Xiao, X.-P., Zhang,
- H.-L., 2016. Yield variation of double-rice in response to climate change in
  Southern China. Eur. J. Agron. 81, 161–168.
- Nisbet, E.G., Manning, M.R., Dlugokencky, E.J., Fisher, R.E., Lowry, D., Michel,
- S.E., Myhre, C.L., Platt, S.M., Allen, G., Bousquet, P., 2019. Very strong
  atmospheric methane growth in the 4 years 2014–2017: Implications for the Paris
  Agreement. Global Biogeochem. Cycles 33, 318–342.
- Pandey, D., Agrawal, M., Bohra, J.S., 2012. Greenhouse gas emissions from rice crop
  with different tillage permutations in rice–wheat system. Agric. Ecosyst. Environ.
  159, 133–144.
- Parton, W.J., Hartman, M., Ojima, D., Schimel, D., 1998. DAYCENT and its land
  surface submodel: description and testing. Glob. Planet. Change 19, 35–48.
- 638 Saunois, M., Stavert, A.R., Poulter, B., Bousquet, P., Canadell, J.G., Jackson, R.B.,
- Raymond, P.A., Dlugokencky, E.J., Houweling, S., Patra, P.K., 2020. The global
  methane budget 2000–2017. Earth Syst. Sci. data 12, 1561–1623.
- 641 Shukla, P.R., Skeg, J., Buendia, E.C., Masson-Delmotte, V., Pörtner, H.-O., Roberts,
- D.C., Zhai, P., Slade, R., Connors, S., Van Diemen, S., 2019. Climate Change
  and Land: an IPCC special report on climate change, desertification, land
  degradation, sustainable land management, food security, and greenhouse gas
  fluxes in terrestrial ecosystems.

646 Smith, J., Smith, P., 2007. Environmental modelling: an introduction. Oxford
647 University Press.

#### 648 Smith, P., Smith, J.U., Powlson, D.S., McGill, W.B., Arah, J.R.M., Chertov, O.G.,

- Coleman, K., Franko, U., Frolking, S., Jenkinson, D.S., 1997. A comparison of
  the performance of nine soil organic matter models using datasets from seven
  long-term experiments. Geoderma 81, 153–225.
- Stehfest, E., Heistermann, M., Priess, J.A., Ojima, D.S., Alcamo, J., 2007. Simulation
  of global crop production with the ecosystem model DayCent. Ecol. Modell. 209,
  203–219.
- Team, R.D.C., 2008. R: A language and environment for statistical computing. R
  Foundation for Statistical Computing, Vienna, Austria. URL http://www.
  Rproject. org.
- United Nations Environment Programme (UNEP) and Climate and Clean Air
  Coalition (CCAC), 2021. Global Methane assessment: Benefits and Costs of
  Mitigating Methane Emissions.
- 661 Wang, H., Tang, S., Han, S., Li, M., Cheng, W., Bu, R., Wang, Y., Cao, W., Wu, J.,
- 2021. Rational utilization of Chinese milk vetch improves soil fertility, rice
  production, and fertilizer use efficiency in double-rice cropping system in East
  China. Soil Sci. Plant Nutr. 1–9.
- Wang, Z., Zhang, X., Liu, L., Wang, S., Zhao, L., Wu, X., Zhang, W., Huang, X.,
- 666 2021. Estimates of methane emissions from Chinese rice fields using the DNDC
  667 model. Agric. For. Meteorol. 303, 108368.
- Watanabe, A., Kimura, M., 1998. Effect of rice straw application on CH4 emission
  from paddy fields: IV. Influence of rice straw incorporated during the previous
  cropping period. Soil Sci. plant Nutr. 44, 507–512.

671	Yang, J.M., Yang, J.Y., Liu, S., Hoogenboom, G., 2014. An evaluation of the
672	statistical methods for testing the performance of crop models with observed data.
673	Agric. Syst. 127, 81–89.

- Yang, Y., Huang, Q., Yu, H., Song, K., Ma, J., Xu, H., 2018. Winter tillage with the
  incorporation of stubble reduces the net global warming potential and
  greenhouse gas intensity of double-cropping rice fi elds. Soil Tillage Res. 183,
  19–27.
- Yearbook, 2014. China statistical yearbook/Agriculture 2014 [WWW Document].
  China Stat. Press. URL http://www.stats.gov.cn/tjsj/ndsj/2014/indexeh.htm
- Zhang, G., Yu, H., Fan, X., Liu, G., Ma, J., Xu, H., 2015. Effect of rice straw

application on stable carbon isotopes, methanogenic pathway, and fraction of

- 682 CH 4 oxidized in a continuously fl ooded rice fi eld in winter season. Soil Biol.
  683 Biochem. J. 84, 75–82.
- Zhang, G., Yu, H., Fan, X., Yang, Y., Ma, J., Xu, H., 2016. Drainage and tillage
  practices in the winter fallow season mitigate CH4 and N2O emissions from a
  double-rice field in China. Atmos Chem Phys 16, 11853–11866.
- Zhang, W., Yu, Y., Huang, Y., Li, T., Wang, P., 2011. Modeling methane emissions
  from irrigated rice cultivation in China from 1960 to 2050. Glob. Chang. Biol. 17,
  3511–3523.
- Zhang, X., Bi, J., Sun, H., Zhang, J., Zhou, S., 2019. Greenhouse gas mitigation
  potential under different rice-crop rotation systems: from site experiment to
  model evaluation. Clean Technol. Environ. Policy 21, 1587–1601.
- Zhang, Y., Li, C., Trettin, C.C., Li, H., Sun, G., 2002. An integrated model of soil,
  hydrology, and vegetation for carbon dynamics in wetland ecosystems. Global
  Biogeochem. Cycles 16, 9-1-9–17.

696	Zhao, Z.,	Cao, L	., Deng,	J.,	Sha,	Ζ.,	Chu,	С.,	Zhou,	D.,	Wu,	S.,	Lv,	W.,	2020.
-----	-----------	--------	----------	-----	------	-----	------	-----	-------	-----	-----	-----	-----	-----	-------

- Modeling CH4 and N2O emission patterns and mitigation potential from paddy
- fields in Shanghai, China with the DNDC model. Agric. Syst. 178, 102743.

# 701 Table

**Table 1** Schedule of field management practices in the experimental plots over the six years from November 20108 to November 2014.

Season	Field Management	2008-2009	2009-2010	2010-2011	2011-2012	2012-2013	2013-2014
Winter							
fallow	Winter tillage	8 Nov 2008	13 Nov 2009	2 Dec 2010	3 Nov 2011	5 Dec 2012	11 Nov 2013
	a	12 4 2000	17 4 2010	10.1. 0011	22.4	<b>2</b> 0 <b>1 2</b> 012	10.1. 2011
Early-rice	Spring tillage	12 Apr 2009	17 Apr 2010	19 Apr 2011	23 Apr 2012	20 Apr 2013	10 Apr 2014
	First flooding	13 Apr 2009	17 Apr 2010	21 Apr 2011	23 Apr 2012	22 Apr 2013	10 Apr 2014
	Basal fertilizers	17 Apr 2009	26 Apr 2010	22 Apr 2011	27 Apr 2012	24 Apr 2013	13 Apr 2014
	Rice translates	17 Apr 2009	27 Apr 2010	23 Apr 2011	27 Apr 2012	24 Apr 2013	13 Apr 2014
	Tillering fertilizers	26 Apr 2009	11 May 2010	14 May 2011	15 May 2012	17 May 2013	29 Apr 2014
	Midseason drainage	8 May 2009~15 May	23 May 2010~27	23 May 2011~31	25 May 2012~5	28 May 2013~3 Jun	22 May 2014~29
		2009	May 2010	May 2011	Jun 2012	2013	May 2014
	Second flooding	16 May 2009~2 Jun	28 May 2010~2 Jun	1 Jun 2011~24 Jun	6 Jun 2012~18 Jun		30 May 2014~16 Jun
		2009	2010	2011	2012	-	2014
	Panicle initiation						
	fertilizers	26 May 2009	12 Jun 2010	16 Jun 2011	12 Jun 2012	14 Jun 2013	10 Jun 2014
	Dry/wet alternation	3 Jun 2009~3 Jul	22 Jun 2010~15 Jul	25 Jun 2011~3 Jul	19 Jun 2012~23		17 Jun 2014~29 Jun
		2009	2010	2011	Jun 2012	-	2014
	Final drainage	4 Jul 2009	16 Jul 2010	4 Jul 2011	24 Jun 2012	4 Jul 2013	30 Jun 2014
	Rice harvest	9 Jul 2009	22 Jul 2010	11 Jul 2011	13 Jul 2012	18 Jul 2013	16 Jul 2014
Late-rice	Tillage	10 Jul 2009	31 Jul 2010	11 Jul 2011	14 Jul 2012	22 Jul 2013	19 Jul 2014
	First flooding	12 Jul 2009	31 Jul 2010	12 Jul 2011	15 Jul 2012	24 Jul 2013	20 Jul 2014
	Basal fertilizers	14 Jul 2009	5 Aug 2010	16 Jul 2011	27 Jul 2012	24 Jul 2013	22 Jul 2014

Rice translates	15 Jul 2009	5 Aug 2010	16 Jul 2011	27 Jul 2012	24 Jul 2013	22 Jul 2014
Tillering fertilizers	29 Jul 2009	23 Aug 2010	3 Aug 2011	14 Aug 2012	13 Aug 2013	4 Aug 2014
Midseason drainage	16 Aug 2009~23	4 Sep 2010~8 Sep	17 Aug 2011~23	22 Aug 2012~1	23 Aug 2013~4 Sep	
	Aug 2009	2010	Aug 2011	Sep 2012	2013	-
Second flooding	24 Aug 2009~6 Sep	8 Sep 2010~29 Sep	24 Aug 2011~4 Sep	2 Sep 2012~1 Oct	5 Sep 2013~19 Sep	1 Sep 2014~22 Sep
	2009	2010	2011	2012	2013	2014
Panicle initiation						
fertilizers	30 Aug 2009	20 Sep 2010	23 Aug 2011	4 Sep 2012	4 Sep 2013	4 Sep 2014
Dry/wet alternation	7 Sep 2009~9 Oct	30 Sep 2010~29 Oct	5 Sep 2011~7 Oct	3 Oct 2012~25 Oct	20 Sep 2013~17	23 Sep 2014~15 Oct
	2009	2010	2011	2012	Oct 2013	2014
Final drainage	10 Oct 2009	30 Oct 2010	8 Oct 2011	26 Oct 2012	18 Oct 2013	16 Oct 2014
Rice harvest	30 Oct 2009	1 Dec 2010	2 Nov 2011	4 Dec 2012	10 Nov 2013	6 Nov 2014

Name of the file	Parameter	Description	Unit	Value
Crop.100	PRDX	Coefficient for calculating potential aboveground monthly	Scaling factor, (g C production)	3.00
		production as a function of solar radiation outside the atmosphere	m <sup>-2</sup> month <sup>-1</sup> Langley <sup>-1</sup>	
	PPDF (1)	Optimum temperature for production for parameterization of a Poisson	°C	25
		Density Function curve to simulate temperature effect on growth		
	PPDF (2)	Maximum temperature for production for parameterization of a Poisson	°C	
		Density Function curve to simulate temperature effect on growth		45
Sitepar.100	CO <sub>2</sub> _to_CH <sub>4</sub>	Fraction of CO <sub>2</sub> from soil respiration used to produce CH <sub>4</sub>		0.15

713	Table 2 The plant production and cultiva	tion parameter files used to calibrate DAYCEN	T model for simulating CH <sub>4</sub> emission and grain yield.
		F F F F F F F F F F F F F F F F F F F	

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Cropping season/parameter	Grain	Leaf	Stem	Root
Early rice (Zhongzao 33)				
Maximum biomass production (kg C ha <sup>-1</sup> y <sup>-1</sup> )	8500	4829	4636	1352
Biomass fraction	0.44	0.25	0.24	0.07
Biomass C/N ratio	51	85	85	30
Thermal degree days	2000			
Water demand (g water/g DM)	508			
Optimum temperature (°C)	25			
Late rice (Nongxiang 98)				
Maximum biomass production (kg C ha <sup>-1</sup> y <sup>-1</sup> )	8500	4829	4636	1352
Biomass fraction	0.44	0.25	0.24	0.07
Biomass C/N ratio	50	85	85	30
Thermal degree days	2850			
Water demand (g water/g DM)	508			
Optimum temperature (°C)	25			

# **Table 3** The crop parameters used to calibrate DNDC model for simulating CH<sub>4</sub> emission and grain yield.

	Annual CH <sub>4</sub> flux (kg C ha <sup>-1</sup> vr <sup>-1</sup> )										
	Treatments	Measured	DAYCENT	RD (%)	DNDC	RD (%)					
727	tillage (W). RD means relative deviation	n between simu	lated and meas	sured emiss	ion/yield.						
726	ha-1), yield-scaled GHG emission (kg CO2	e-eq ha <sup>-1</sup> yr <sup>-1</sup> ) by	the treatment o	f stubble ir	ncorporatio	n (S), winter till	ge with stub	oble incorpo	oration (WS)	and winte	er
725	Table 4. Comparison between the DAYC	ENT- and DNI	DC-simulated a	nd measur	ed average	e annual CH <sub>4</sub> (kg	C ha <sup>-1</sup> yr <sup>-1</sup> )	fluxes, earl	ly- and late-r	rice yield	(t

727	tillage (W).	RD 1	neans rel	ative	dev	iation	between	simulated	and	measured	emission/	yield.
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Annual CH flux $(lra C ha^{-1} vv^{-1})$											
Annual CH <sub>4</sub> flux (kg C ha <sup>-1</sup> yr <sup>-1</sup> )											
S	175±26A <sup>a</sup>	173±15A	1	173±15A	1						
WS	152±29AB	148±19AB	3	153±20A	-1						
W	111±26B	138±18B	-24	117±22B	-5						
Early rice yield (t ha <sup>-1</sup> )											
S	6.3±0.2A	6.4±0.4A	-1	6.1±0.3A	2						
WS	6.6±0.3A	6.4±0.4A	3	6.7±0.6A	-2						
W	6.5±0.2A	6.4±0.4A	2	7.0±0.8A	-8						
Late rice yields (t ha <sup>-1</sup> )											
S	$6.4 \pm 0.8 A$	6.3±0.8A	1	6.6±0.7A	-3						
WS	6.5±0.9A	6.3±0.8A	4	6.8±0.7A	-4						
W	6.3±0.9A	6.3±0.8A	0	6.7±0.7A	-6						
Yield-scaled GHG <sub>CH4</sub> (kg CO <sub>2</sub> -eq ha <sup>-1</sup> yr <sup>-1</sup> )											
S	0.52±0.03A	0.52±0.03A	0	0.51±0.03A	2						
WS	0.43±0.03AB	0.44±0.03AB	-2	0.43±0.03AB	0						
W	0.32±0.03B	0.41±0.03B	-28	0.32±0.03B	0						

<sup>a</sup> Values followed by the same letter are not significantly different within the treatments at p < 0.05 based on Turkey tests. 

731 **Table 5.** Statistical describing the performance of the DAYCENT and DNDC models for the simulations of daily CH<sub>4</sub> fluxes under different treatments in the

Treatment	Model	Measured (kg C ha <sup>-1</sup> )	RMSE (kg C ha <sup>-1</sup> )	rRMSE (%)	EF	r	M (kg C ha <sup>-1</sup> )
S (n=398)	DAYCENT	0.67	0.85	127	0.28	0.60***	0.04 <sup>ns</sup>
	DNDC	0.07	0.85	128	0.27	0.61***	0.03 <sup>ns</sup>
WC (~ 200)	DAYCENT	0.58	0.75	129	0.22	0.58***	0.04 <sup>ns</sup>
WS (II-398)	DNDC		0.72	124	0.28	0.63***	0.02 <sup>ns</sup>
W (n=335)	DAYCENT	0.42	0.63	150	-0.07	0.59***	-0.08*
	DNDC		0.52	140	0.08	0.60***	-0.02 <sup>ns</sup>

double rice paddy from November 2008 to November 2014. The n is the number of measured CH<sub>4</sub> fluxes from November 2008 to November 2014.

<sup>a</sup>S, stubble incorporation without winter tillage; WS, winter tillage with stubble incorporation; W, winter tillage without stubble incorporation.

\* Significant correlation (r) between modelled and measured values at p < 0.05, or significance mean error (M) at p=0.025.

\*\*\* Significant correlation (r) between modelled and measured values at p < 0.001.

<sup>ns</sup> Non-significant between modelled and measured values at p < 0.05, or no significance mean error (M) at p=0.025.

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#### 741 Figure legends

**Fig. 1.** Sensitivity of CH<sub>4</sub> fluxes and yield to changes in the input parameters. SOC: soil organic carbon content (from 0.5 to 1.5 times the baseline). pH: soil pH (from "baseline – 1" to "baseline + 1"). N fer: application of N fertilizer (from 0.5 to 1.5 times the baseline). T: air temperature (from "baseline – 2" to "baseline + 2"). The SOC, pH, N fertilizer and daily average air temperature were 0.016 g kg<sup>-1</sup>, 4.6, 360 kg N ha<sup>-1</sup> and 18.16 °C.

**Fig. 2.** Comparison between the DAYCENT- and DNDC-simulated and measured daily CH<sub>4</sub> flux (Kg C ha<sup>-1</sup> d<sup>-1</sup>) from November 2008 to

November 2014 for the treatments of stubble incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage (W).

**Fig. 3.** Measured annual CH<sub>4</sub> emission (kg C ha<sup>-1</sup> yr<sup>-1</sup>) and yield (t ha<sup>-1</sup> yr<sup>-1</sup>) over the six annual rotation cycles from November 2008 to

November 2014 for the treatments of stubble incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage (W). Values are

the means with standards deviations shown by vertical bars (n=3); uppercase letters indicate significant differences within years at p < 0.05.

**Fig. 4.** Comparison between DAYCENT- and DNDC-simulated and measured seasonal CH<sub>4</sub> (kg C ha<sup>-1</sup>) for three treatments of stubble

incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage (W) from November 2008 to November 2014.

Fig. 5. Relationship between the DAYCENT- and DNDC-simulated and measured yields of early and late-paddy rice for the treatments of

stubble incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage (W).

1	Supporting Information for
2	
3	Modelling methane emissions and grain yields for a double-rice system in
4	Southern China with DAYCENT and DNDC
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13	
14	<sup>1</sup> Contributed equally to this work
15	* Corresponding author. E-mail address: guo.yang.2021@outlook.com
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17	
18	Summary:
19	Number of Figures: 1
20	Number of Tables: 1



**Fig. S1** Measured daily air temperature (°C) and rainfall (cm) during the experimental period (from November 2008 to November 2014) in the Yingtan of China.

Year/Month	Rainfall (cm)						Mean a	Mean air temperature (°C)								
	2008	2009	2010	2011	2012	2013	2014	Mean	2008	2009	2010	2011	2012	2013	2014	Mean
January	10	3.8	11.3	5.6	13.0	3.2	1.9	6.5	3.9	4.6	6.6	2.9	5.3	5.1	6.7	5.2
February	7.0	6.8	15.3	5.0	9.7	16.7	16.6	11.7	5.3	11.5	10.1	8.4	6.3	9.1	7.3	8.8
March	17.4	29.8	22.8	8.8	29.9	27.5	28.3	24.5	13.6	12.6	12.5	11.2	11.7	13.9	13.4	12.5
April	25.8	19.0	34.5	15.2	25.2	20.4	16.1	21.7	19.0	19.1	15.5	19.1	19.3	16.9	19.3	18.2
May	14.2	18.5	41.5	11.9	36.2	25.2	21.5	25.8	23.8	23.0	22.7	22.5	23.3	23.4	22.3	22.9
June	26.1	10.3	58.3	49.4	38.3	39.8	22.8	36.5	25.8	27.1	24.5	26.1	26.0	26.4	25.8	26.0
July	18.9	12.1	22.2	11.8	25.1	2.8	22.6	16.1	29.7	29.2	29.6	29.9	29.8	30.1	28.8	29.6
August	5.9	14.0	7.4	19.2	36.2	3.3	15.4	15.9	29.1	29.0	29.9	28.8	28.5	30.6	27.5	29.1
September	5.2	2.3	14.0	2.0	14.4	3.4	4.1	6.7	26.7	26.9	26.6	25.2	23.3	25.2	26.5	25.6
October	6.1	1.3	7.6	6.5	3.5	1.8	0.3	3.5	20.6	20.8	18.4	19.3	19.6	19.3	20.6	19.6
November	13.0	13.2	18.4	7.1	25.5	11.5	6.7	13.7	12.5	10.8	13.4	16.5	11.9	13.2	14.4	13.4
December	2.1	6.7	20.3	4.2	13.2	7.6	9.5	10.2	7.2	6.9	7.9	6.7	6.5	5.9	6.0	6.6

 Table S1 Monthly mean air temperature and rainfall from 2008 and 2014 at the field site.

# **Table**

**Table 1** Schedule of field management practices in the experimental plots over the six years from November 20108 to November 2014.

Season	Field Management	2008-2009	2009-2010	2010-2011	2011-2012	2012-2013	2013-2014
Winter							
fallow	Winter tillage	8 Nov 2008	13 Nov 2009	2 Dec 2010	3 Nov 2011	5 Dec 2012	11 Nov 2013
Early-rice	Spring tillage	12 Apr 2009	17 Apr 2010	19 Apr 2011	23 Apr 2012	20 Apr 2013	10 Apr 2014
	First flooding	13 Apr 2009	17 Apr 2010	21 Apr 2011	23 Apr 2012	22 Apr 2013	10 Apr 2014
	Basal fertilizers	17 Apr 2009	26 Apr 2010	22 Apr 2011	27 Apr 2012	24 Apr 2013	13 Apr 2014
	Rice translates	17 Apr 2009	27 Apr 2010	23 Apr 2011	27 Apr 2012	24 Apr 2013	13 Apr 2014
	Tillering fertilizers	26 Apr 2009	11 May 2010	14 May 2011	15 May 2012	17 May 2013	29 Apr 2014
	Midseason drainage	8 May 2009~15	23 May 2010~27	23 May 2011~31	25 May 2012~5	28 May 2013~3	22 May 2014~29
		May 2009	May 2010	May 2011	Jun 2012	Jun 2013	May 2014
	Second flooding	16 May 2009~2 Jun	28 May 2010~2 Jun	1 Jun 2011~24 Jun	6 Jun 2012~18 Jun		30 May 2014~16
		2009	2010	2011	2012	-	Jun 2014
	Panicle initiation						
	fertilizers	26 May 2009	12 Jun 2010	16 Jun 2011	12 Jun 2012	14 Jun 2013	10 Jun 2014
	Dry/wet alternation	3 Jun 2009~3 Jul	22 Jun 2010~15 Jul	25 Jun 2011~3 Jul	19 Jun 2012~23		17 Jun 2014~29 Jun
		2009	2010	2011	Jun 2012	-	2014
	Final drainage	4 Jul 2009	16 Jul 2010	4 Jul 2011	24 Jun 2012	4 Jul 2013	30 Jun 2014
	Rice harvest	9 Jul 2009	22 Jul 2010	11 Jul 2011	13 Jul 2012	18 Jul 2013	16 Jul 2014
Late-rice	Tillage	10 Jul 2009	31 Jul 2010	11 Jul 2011	14 Jul 2012	22 Jul 2013	19 Jul 2014
	First flooding	12 Jul 2009	31 Jul 2010	12 Jul 2011	15 Jul 2012	24 Jul 2013	20 Jul 2014
	Basal fertilizers	14 Jul 2009	5 Aug 2010	16 Jul 2011	27 Jul 2012	24 Jul 2013	22 Jul 2014

Rice translates	15 Jul 2009	5 Aug 2010	16 Jul 2011	27 Jul 2012	24 Jul 2013	22 Jul 2014
Tillering fertilizers	29 Jul 2009	23 Aug 2010	3 Aug 2011	14 Aug 2012	13 Aug 2013	4 Aug 2014
Midseason drainage	16 Aug 2009~23	4 Sep 2010~8 Sep	17 Aug 2011~23	22 Aug 2012~1	23 Aug 2013~4	
	Aug 2009	2010	Aug 2011	Sep 2012	Sep 2013	-
Second flooding	24 Aug 2009~6 Sep	8 Sep 2010~29 Sep	24 Aug 2011~4 Sep	2 Sep 2012~1 Oct	5 Sep 2013~19 Sep	1 Sep 2014~22 Sep
	2009	2010	2011	2012	2013	2014
Panicle initiation						
fertilizers	30 Aug 2009	20 Sep 2010	23 Aug 2011	4 Sep 2012	4 Sep 2013	4 Sep 2014
Dry/wet alternation	7 Sep 2009~9 Oct	30 Sep 2010~29 Oct	5 Sep 2011~7 Oct	3 Oct 2012~25 Oct	20 Sep 2013~17	23 Sep 2014~15 Oct
	2009	2010	2011	2012	Oct 2013	2014
Final drainage	10 Oct 2000	20 Oct 2010	8 Oct 2011	26 Oct 2012	18 Oct 2013	16 Oct 2014
-	10 Oct 2009	30 Oct 2010	8 001 2011	20 000 2012	10 000 2015	10 000 2014

	Name of the file	Parameter	Description	Unit	Value
	Crop.100	PRDX	Coefficient for calculating potential aboveground monthly	Scaling factor, (g C	3.00
			production as a function of solar radiation outside the atmosphere	production) $m^{-2}$ month <sup>-1</sup>	
				Langley <sup>-1</sup>	
		PPDF (1)	Optimum temperature for production for parameterization of a Poisson	°C	25
			Density Function curve to simulate temperature effect on growth		
		PPDF (2)	Maximum temperature for production for parameterization of a Poisson	°C	
			Density Function curve to simulate temperature effect on growth		45
	Sitepar.100	CO <sub>2</sub> _to_CH <sub>4</sub>	Fraction of CO <sub>2</sub> from soil respiration used to produce CH <sub>4</sub>		0.15
14					
15					
16					
17					
18					

13	Table 2 The plant production and cultivation parameter files used to calibrate DAYCENT model for simulating CH4 emission and grain yield.

Cropping season/parameter	Grain	Leaf	Stem	Root
Early rice (Zhongzao 33)				
Maximum biomass production (kg C ha <sup>-1</sup> y <sup>-1</sup> )	8500	4829	4636	1352
Biomass fraction	0.44	0.25	0.24	0.07
Biomass C/N ratio	51	85	85	30
Thermal degree days	2000			
Water demand (g water/g DM)	508			
Optimum temperature (°C)	25			
Late rice (Nongxiang 98)				
Maximum biomass production (kg C ha <sup>-1</sup> y <sup>-1</sup> )	8500	4829	4636	1352
Biomass fraction	0.44	0.25	0.24	0.07
Biomass C/N ratio	50	85	85	30
Thermal degree days	2850			
Water demand (g water/g DM)	508			
Optimum temperature (°C)	25			

Table 3 The crop parameters used to calibrate DNDC model for simulating CH<sub>4</sub> emission and grain yield.

25 **Table 4.** Comparison between the DAYCENT- and DNDC-simulated and measured average annual CH<sub>4</sub> (kg C ha<sup>-1</sup> yr<sup>-1</sup>) fluxes, early- and late-rice yield (t ha<sup>-1</sup>

<sup>1</sup>), yield-scaled GHG emission (kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>) by the treatment of stubble incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage

Treatments	Measured	DAYCENT	RD (%)	DNDC	RD (%)
Annual CH4 flux (kg C ha <sup>-1</sup> yr <sup>-1</sup> )					
S	175±26A <sup>a</sup>	173±15A	1	173±15A	1
WS	152±29AB	148±19AB	3	153±20A	-1
W	111±26B	138±18B	-24	117±22B	-5
Early rice yield (t ha <sup>-1</sup> )					
S	6.3±0.2A	6.4±0.4A	-1	6.1±0.3A	2
WS	6.6±0.3A	6.4±0.4A	3	6.7±0.6A	-2
W	6.5±0.2A	6.4±0.4A	2	7.0±0.8A	-8
Late rice yields (t ha <sup>-1</sup> )					
S	$6.4 \pm 0.8 A$	6.3±0.8A	1	6.6±0.7A	-3
WS	6.5±0.9A	6.3±0.8A	4	6.8±0.7A	-4
W	6.3±0.9A	6.3±0.8A	0	6.7±0.7A	-6
Yield-scaled GHG <sub>CH4</sub> (kg CO <sub>2</sub> -eq ha <sup>-1</sup> yr <sup>-1</sup> )					
S	0.52±0.03A	0.52±0.03A	0	0.51±0.03A	2
WS	0.43±0.03AB	$0.44 \pm 0.03 AB$	-2	0.43±0.03AB	0
W	$0.32{\pm}0.03B$	$0.41 \pm 0.03 B$	-28	$0.32 \pm 0.03 B$	0

27 (W). RD means relative deviation between simulated and measured emission/yield.

<sup>a</sup> Values followed by the same letter are not significantly different within the treatments at p < 0.05 based on Turkey tests.

**Table 5.** Statistical describing the performance of the DAYCENT and DNDC models for the simulations of daily CH<sub>4</sub> fluxes under different treatments in the

Treatment	Model	Measured (kg C ha <sup>-1</sup> )	RMSE (kg C ha <sup>-1</sup> )	rRMSE (%)	EF	r	M (kg C ha <sup>-1</sup> )
S (n=398)	DAYCENT	0.47	0.85	127	0.28	0.60***	0.04 <sup>ns</sup>
	DNDC	0.07	0.85	128	0.27	0.61***	0.03 <sup>ns</sup>
WS (n=398)	DAYCENT	0.58	0.75	129	0.22	0.58***	0.04 <sup>ns</sup>
	DNDC		0.72	124	0.28	0.63***	0.02 <sup>ns</sup>
W (n=335)	DAYCENT	0.42	0.63	150	-0.07	0.59***	-0.08*
	DNDC		0.52	140	0.08	0.60***	-0.02 <sup>ns</sup>

double rice paddy from November 2008 to November 2014. The n is the number of measured CH<sub>4</sub> fluxes from November 2008 to November 2014.

<sup>33</sup> <sup>a</sup>S, stubble incorporation without winter tillage; WS, winter tillage with stubble incorporation; W, winter tillage without stubble incorporation.

\* Significant correlation (r) between modelled and measured values at p < 0.05, or significance mean error (M) at p=0.025.

\*\*\* Significant correlation (r) between modelled and measured values at p < 0.001.

 $^{ns}$  Non-significant between modelled and measured values at p < 0.05, or no significance mean error (M) at p=0.025.

#### 39 Figure legends



40

41 Fig. 1. Sensitivity of CH<sub>4</sub> fluxes and yield to changes in the input parameters. SOC: soil organic carbon content (from 0.5 to 1.5 times the baseline).

42 pH: soil pH (from "baseline – 1" to "baseline + 1"). N fer: application of N fertilizer (from 0.5 to 1.5 times the baseline). T: air temperature (from 43 "baseline – 2" to "baseline + 2"). The SOC, pH, N fertilizer and daily average air temperature were 0.016 g kg<sup>-1</sup>, 4.6, 360 kg N ha<sup>-1</sup> and 18.16 °C.



Fig. 2. Comparison between the DAYCENT- and DNDC-simulated and measured daily  $CH_4$  flux (Kg C ha<sup>-1</sup> d<sup>-1</sup>) from November 2008 to November





Fig. 3. Measured annual CH4 emission (kg C ha-1 yr-1) and yield (t ha-1 yr-1) over the six annual rotation cycles from November 2008 to November 2014 for the treatments of stubble incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage (W). Values are

50 the means with standards deviations shown by vertical bars (n=3); uppercase letters indicate significant differences within years at p < 0.05.



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Fig. 4. Comparison between DAYCENT- and DNDC-simulated and measured seasonal
CH<sub>4</sub> (kg C ha<sup>-1</sup>) for three treatments of stubble incorporation (S), winter tillage with
stubble incorporation (WS) and winter tillage (W) from November 2008 to November
2014.



Fig. 5. Relationship between the DAYCENT- and DNDC-simulated and measured yields of early and late-paddy rice for the treatments of stubble
 incorporation (S), winter tillage with stubble incorporation (WS) and winter tillage (W).