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### 2 Distribution and thermal niche of the common skate species complex in the

**North-East Atlantic** 

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14 Running page header: Common skate distribution and thermal niche

ABSTRACT: Temperature is one of the most significant variables affecting the geographic 16 distribution and physiology of elasmobranchs. Differing thermal gradients across a species' 17 range can lead to adaptive divergence and differing developmental times, an important 18 consideration for recruitment rates of exploited species. The critically endangered common 19 skate (formerly *Dipturus batis*) has been divided into two species, the flapper skate (D. 20 intermedius) and blue skate (D. batis), both of which have undergone dramatic population 21 declines. Here we examine the environmental thermal and geographic distribution of these 22 23 species, using observations from scientific trawling surveys and recreational angling around 24 the British Isles. As similar sized specimens of the two species can be confused, we validated 25 species identity using molecular genetic techniques. Both species had more extensive geographic ranges than previously reported and different spatial patterns of abundance. The 26 27 distribution of the blue skate appears to reflect its partiality to thermally less variable and warmer waters, while flapper skate were found in more variable and notably colder areas. The 28

- 29 thermal range and current geographic distribution of these species indicate future projected
- 30 climate change could have a differential impact on distribution of flapper and blue skate in the
- 31 North-East Atlantic.
- 32 Key words: Distribution  $\cdot$  thermal niche  $\cdot$  common skate species complex  $\cdot$  Dipturus  $\cdot$  flapper
- 33 skate · blue skate · molecular markers

#### 1. INTRODUCTION

Temperature is a fundamental environmental variable affecting the distribution and life-history of marine organisms (Wallman & Bennett 2006). It is a key determinant in the physiological performance of poikilotherms, and some species have thermoregulatory behaviors that drive habitat choice (Vaudo & Heithaus 2013). Temperature regulates feeding, growth, survival and development in fish (Pistevos et al. 2017), and is an important factor underlying the distribution and depth preferences of fish stocks (Perry et al. 2005, Dulvy et al. 2008, Rutterford et al. 2015). Temperature is known to affect the metabolism, development, growth, movement patterns and reproduction of elasmobranchs (Sinclair et al. 2016) and their overall life-history (Schlaff et al. 2014). Ambient temperature is particularly important to oviparous species because of the inability of developing embryos to utilize or avoid changes in the surrounding environment by engaging in thermotaxic behavior (Di Santo 2015). Embryos exposed to differing temperatures within the egg case have plastic growth rates that result in variable developmental durations (Pretorius & Griffiths 2013) across their distribution range.

Knowledge of the geographic range and habitat preferences of fish is fundamental for effective management, which can facilitate measures to limit the impact of anthropogenic threats, such as the establishment of marine protected areas (MPAs) (Tserpes et al. 2013). The spatial distribution and migration patterns of many elasmobranchs in the waters of the British Isles has only recently begun to be revealed with advances in electronic tagging (Hunter et al. 2005, Saunders et al. 2011, Neat et al. 2014, Doherty et al. 2017a,b, Biais, et al. 2017). Efforts to compile survey data have also been valuable for clarifying fish distributions (Heeson et al. 2015) in the Celtic Sea, North Sea, and Baltic Sea. Thermal ranges, however, remain difficult to assess and interpret because many elasmobranch populations are in decline due to

overfishing, with some species already extirpated from areas within their former geographic range (Brander 1981, Ellis et al. 2005, Hunter et al. 2005). The situation is further exacerbated by the lack of species-specific landing data from commercial fisheries (Stevens et al. 2000), which can mask the extent of decline and collapse.

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The common skate (formerly Dipturus batis) was once frequently encountered throughout European waters but has undergone extreme population declines and range contractions due to fishing practices over the last century (Walker & Hislop 1998). As a result of this decline the species was classified as critically endangered by the IUCN in 2006. Soon after this it became apparent that it was actually two species that could be differentiated on genetic and morphological characters (Griffiths et al. 2010, Iglesias et al. 2010). The most recent revision of their nomenclature (Last et al. 2016) defined the larger-bodied flapper skate as *Dipturus* intermedius (formerly D. intermedia) and the smaller-bodied blue skate as Dipturus batis (formerly *D. flossada*). The flapper skate reaches lengths over 2.5 m and attains sexual maturity at ~19 years of age, while the smaller blue skate reaches lengths of ~1.4 m and sexual maturity at ~11 years of age (Iglesias et al. 2010). The species complex is found off the Scottish west coast, the Celtic Sea, Rockall Bank, Iceland, and rarely encountered in the North Sea and Irish Sea (Dulvy et al. 2006), however, the spatial distribution and extent of overlap of the two species is not well resolved. Griffiths et al. (2010) showed some evidence of spatial segregation linked to thermal range, but their analysis was based on sea surface temperature, which does not reflect the true thermal regime experienced by benthic skates that spend most of their time living on or near the seabed. Both species can be found in a variety of habitats and range of depths from the surface and coast to the continental slope and depths up to 500 m (Wearmouth & Sims 2009, Griffiths et al. 2010, Neat et al. 2014, Bendall et al. 2017) and the full geographic

range extends to Iceland and Norway in the north and into Bay of Biscay and the Mediterranean in the South.

Using data from trawl surveys, recreational angler catches and modelled bottom temperature, we test the null hypothesis that there is no difference in the spatial distribution and thermal ranges of the two species. We predicted that the spatial distribution of the common skate complex reflects differences in thermal preferences between the two species; with the larger flapper skate inhabiting colder more northerly waters and smaller blue skate warmer more southern and offshore areas (Griffiths et al. 2010). Molecular markers were used to verify specimens and we present data resolving the spatial distribution of flapper and blue skate populations around the British Isles and the offshore Rockall plateau.

#### 2. METHODS

95 2.1. Sampling

Tissue samples from 915 specimens of 'common skate' were obtained between 2008-2013 from 8 different regions (Figure 1A) around the British Isles by Marine Scotland Science, the Centre for Environment Fisheries and Aquaculture Science (CEFAS), and from recreational anglers. Samples were obtained from: the Celtic Sea (CS; n=188), northern Scottish and Irish continental shelf (NSHLF; n=56), southern Scottish and Irish continental shelf (SSHLF; n=48), Rockall Bank (ROCK; n=129), western coast of Scotland (SWC; n=427; includes the Loch Sunart to Sound of Jura MPA), Ireland (IRE; n=7), far north of Scotland in Orkney and Shetland (FN; n=58) and a deep-sea area to the North of Shetland (DS; n=2; Table 1). Geographic coordinates, total length, and sex were recorded for all sampled skate, and depth was recorded during Marine Scotland scientific surveys. Length frequency distributions, and

size comparisons between males and females of both species captured in each location were examined. Sex ratios, in each species, were examined using a 2-tailed binomial test in R (R Core Team 2020) to determine if the proportions of males and females were equal. The maturation status of each skate was assigned, using estimates of length at 50% maturity ( $L_{50}$ ) from Iglesias et al. (2010), and numbers of mature and immature individuals were determined for both species.

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## 2.2. Molecular Species Identification

Morphological characteristics (Iglesias et al. 2010), most notably size of mature specimens, eye colour and dorsal patterning, were used to differentiate blue and flapper skate. Molecular markers were required to validate the assignment of 406 individuals considered morphologically ambiguous by collectors from the CS (n=188), SWC (n=141), IRE (n=7), FN (n=58), NSHLF (10) and DS (n=2) surveys. A small tissue sample was removed from the tail or wing and immediately preserved in 95% ethanol or RNAlater® (Thermo Fisher Scientific) before returning fish to the water. Genomic DNA was extracted from ~20 mg of tissue using a modified phenol-chloroform protocol (Sambrook et al. 1989). Five microsatellites (LERI 21, 33, 34, 44 and 50) from El Nagar et al. (2010), which previously showed clear species delineation (Griffiths et al. 2010, McCutchen 2012), were used to identify species. PCR primers were fluorescently-labelled with PET, NED, HEX and 6-FAM (Applied Biosystems) and fit into a multiplex with a LIZ500 internal size standard. PCRs were performed in an 11 µl reaction that contained 2 µl of 10ng/µl genomic DNA, 3 µl reaction buffer (Bioline), 1 µl forward primer, 1 µl reverse primer, 0.2 µl BIOTAQ DNA Polymerase (Bioline) and 3.8 µl H<sub>2</sub>O on a T-Gradient Cycler (Biometra). Thermocycling conditions included: initial denaturation of 3 min at 94°C; 30 cycles of denaturation at 94°C for 30s;

annealing at 53°C for 30s; extension at 72°C for 30s; a final extension step of 72°C for 10 min. PCR products were separated on an Applied Biosystems 3730 DNA Analyzer at the Tayside Centre for Genomic Analysis (University of Dundee, Dundee, Scotland). Genotypes were called manually using GeneMarker Version 2.2.0. Species membership was assigned using STRUCTURE v2.3.4 (Pritchard et al. 2000), which employs a Bayesian approach to identify the most probable number of clusters produced from the data (K). Ten replicates with 1,000,000 MCMC iterations and a 200,000 burn-in were used, and the number of clusters set between 1 and 6 to ensure other species were not included. STRUCTURE Harvester v0.6.93 (Earl & vonHoldt 2012) was used to examine the statistically best supported value of K, and the results summarized in CLUMPAK (Kopelman et al. 2015). Bayesian clustering indicated that the optimal value of K was 2 and assigned species membership for 406 ambiguously identified individuals (supplementary Figure S1), of which 179 were flapper skate and 227 blue skate. 

### 2.3. Environmental data and species distribution

Data from the ICES hydrographic CTD database were used in a linear regression model (in combination with Kriging) to generate a surface of average monthly bottom temperature across the study area. Interpolation assigned a temperature (at a spatial resolution of 1 ICES rectangle of 0.5 degree X 0.5 degree) for each latitude and longitude position where a skate was sampled. For each sampled location values of minimum and maximum temperature were extracted for each month skate were sampled. Temperature and depth data were tested for normal distributions using a Shapiro-Wilk normality test, visualized in R (R Core Team 2020); however, the data were not normally distributed, and transformation failed to achieve normality. A non-parametric Mann-Whitney U test was used to determine if the observed

differences between the means of minimum temperature, maximum temperature, temperature range and depth were statistically significant (P<0.05) when comparing the two species. To avoid issues of pseudo-replication, only a single skate, or one skate of each species, per haul or sampling event was included in statistical analyses, which consisted of 100 flapper skate and 73 blue skate in temperature analyses. Statistical tests were carried out in R (R Core Team 2020), and boxplots were produced in R (R Core Team 2020) using the package ggplot2 (Wickham 2016). The distribution of skates sampled in this study was mapped using the coordinates of 915 sampling locations. Scientific bottom trawls from Marine Scotland Science surveys were used to determine the numbers of each skate species caught per hour. The midpoint of each trawl was calculated and then summarised on a regular hexagonal grid with a cell width of 20 km by taking the mean value across those hauls contributing to each grid cell. This resulted in a mean catch per unit effort value for each grid cell. Derivation of catch rates was completed in R (R Core Team 2020), and the maps were produced using QGIS version 3.4.10 (2019).

### 3. RESULTS

### 3.1. Length, Maturity and Sex Composition

Flapper skate total lengths ranged from 21 to 230 cm, but 96% of sampled flapper skates were below  $L_{50}$  and only 4% were reproductively mature (>185.5 cm). The most abundant size group was the 41-60 cm length class, comprising 46% of flapper skate, followed by the 61-80 cm length class at 21%. The largest flapper skate were sampled in the Loch Sunart to Sound of Jura MPA (west coast of Scotland), Ireland and the Celtic Sea, while the remaining populations were mostly composed of small juveniles with a few larger individuals present (Figure S2A in

supplementary information). The flapper skate sex ratio was not significantly different

178 (P=0.415) with 234 females and 253 males sampled (Table S1).

Blue skate total lengths ranged from 21 to 148 cm, and over 40% were reproductively mature (>~115 cm). Nearly 45% of blue skates were in the 121-140 cm size class, which was followed by 14% in both the 81-100 cm and 101-120 cm length classes. The largest blue skate were sampled in the Celtic Sea (Figure S2B in supplementary information), with approximately 80% above  $L_{50}$ ; however, only 15% of skate sampled in Rockall were at or above  $L_{50}$ . Only one blue skate from the far north of Scotland was at  $L_{50}$ , and no mature skate were sampled on the west coast of Scotland. The blue skate sex ratio was unequal (P=0.045) with a higher proportion of females (n=200) than males (n=161; Table S1).

## 3.2. Spatial Distribution

Sampling locations were mapped for 915 skate sampled in this study, of which 554 were flapper skate and 361 were blue skate. Both species had wider geographic distributions than previously reported (Griffith et al. 2010). Blue skate were found at latitudes ranging between 49.43°N and 60.58°N and longitudes of 2.32°W to 16.31°W and flapper skate were sampled between 49.48°N and 62.08°N latitude and -0.04°W to -9.53°W longitude (Figure 1B). Flapper skate were found from the Celtic Sea to north of Shetland; however, greatest concentrations of flapper skate were found along the western coast and continental shelf of Scotland (Figure 1B and C). Blue skate also had a wide range from the Celtic Sea to north of Orkney with an extensive representation around Rockall and the Celtic Sea (Figure 1B and D).

### 3.3. Bottom temperature and bathymetry

In this study, flapper skate occur in waters ranging from 4.96°C to 15.50°C, while blue skate were sampled in temperatures between 7.44°C and 13°C (Figure 2). Flapper skate were found in significantly colder minimum temperatures (W=5765; P<0.001; mean=8.41°C) than blue

skate (mean=9.07°C; Figure 2A, Table S1) and significantly warmer maximum temperatures (W=1704; P<0.001; flapper skate mean=12.01°C; blue skate mean=11.09°C; Figure 2B, supplementary Table S1). Flapper skate were found in a significantly (W=1599; P<0.001) wider range of temperatures (0.424 to 7.830; mean=3.59) than blue skate (0.458 to 5.365; mean=2.01; Figure 3, Table S1).

Recorded depth ranges were similar for both species with flapper skate sampled from depths ranging between 51-500 m and blue skate sampled from depths of 56-550 m.

### 4. DISCUSSION

### 4.1 Spatial distribution

This study represents a comprehensive assessment of the geographic and thermal ranges of the common skate complex. Data collected from scientific trawling surveys and recreational angling indicate that flapper and blue skate have overlapping geographical distributions and are more widespread throughout the British Isles than previously reported (Griffiths et al. 2010). Blue and flapper skate appear to cohabit many of the same geographic areas and are often encountered in the same hauls, apart from the offshore Rockall Bank, where only blue skate were recorded. It is notable that the Rockall Bank is isolated from all other areas by water depths in excess of 1500 m and also has only a tiny proportion of its area that is shallower than 100 m. Strong spatial structuring of the two species was evident in this study, similar to that reported by Griffiths et al. (2010). Blue skate appear to predominate in offshore areas, whereas flapper skate are also found closer inshore. There was, however, no evidence of a strong latitudinal or allopatric separation as previously proposed (Griffiths et al. 2010). Blue skate, thought to be the "southern" species, were recorded as far north as Shetland and along the west coast of Scotland, whilst large sub-adult and reproductively mature adult flapper skate, the

"northern" species, were recorded far to the south and in the offshore Celtic Sea. It is important to appreciate that populations of both species have been heavily depleted over the past century or more and that the current distribution may be highly patchy due to local extirpation and small areas of refuge.

Both species appeared to be absent between 50°N and 54°N of latitude in the Irish Sea and along the Irish continental shelf. This, however, reflects much less sampling of these areas. Historically, large common skate were an important component of fisheries in the Irish Sea, but catch rates began to decline drastically in the 1950s until they disappeared altogether in the 1970s (Brander 1981). Reported occurrences remain rare in the Irish Sea with individuals occasionally recorded in remote sites (Iglesias et al. 2010). No blue skate were sampled off the west coast of Ireland during this study, but several large adult and sub-adult flapper skate used in this study were sampled in Irish waters and records from recreational catch and release fisheries (Scottish Shark Tagging Program pers. comm.) suggest flapper skate are encountered in the seas off Northern Ireland.

### 4.2 Bottom temperature and bathymetry

There were significant differences between the mean minimum, maximum and range of temperatures the two species were found at. Although both species occurred in many of the same areas, the results of this study indicate differences in their thermal ranges, most notably with respect to the average minimum and maximum bottom temperatures experienced. In this study, flapper skate were found in cooler and warmer waters than blue skate. This may reflect their preference for thermally and bathymetrically variable inshore habitats that include both the deep, cold sea lochs along the west coast and islands of Scotland (Wearmouth & Sims 2009;

Neat et al. 2014) and close-by shallow coastal areas with highly variable seasonal temperatures. Blue skate, more closely associated with warmer temperatures, were more prevalent in the oceanic areas of Rockall and the Celtic Sea. Although present in the more northerly latitudes they were less abundant. Flapper skate were found in a wider range of temperatures, including the coldest and warmest temperatures recorded in this study, while blue skate appeared to predominate in areas where temperatures are moderated year-round by warm currents, such as the Rockall Bank and Celtic Sea. A major caveat of this type of analysis is that large elasmobranchs are likely to migrate throughout the year, and, therefore, the point of capture may not be representative of their annual thermal experience. However, Neat et al. (2014) showed that a significant proportion of flapper skate in the Loch Sunart to Sound of Jura MPA demonstrated site fidelity for most months of the year. This suggests that the temperatures in our dataset are likely representative of the integrated averages experienced by many individuals within these populations for a substantial fraction of the year.

Temperature can have a substantial effect on the development, survival and metabolic rate of embryonic oviparous elasmobranchs (Pretorius & Griffiths 2013, Di Santo 2015). The incubation period of catshark (*Poroderma pantherinum* and *Haploblepharus pictus*) embryos was shortened by up to 53% and embryos grew up to twice as fast when the temperature of developing eggs was raised by 3°C (Pretorius & Griffiths 2013). The metabolic stability of embryonic little skate (*Leucoraja erinacea*) from two geographic locations declined after reaching the thermal optimum, but the southern population was affected less by increased temperature, suggesting the narrower thermal tolerance of the northern population led to increased metabolic costs at higher temperatures (Di Santo 2015). Although little is known about the embryonic development of flapper or blue skate, fluctuations in temperatures cause an exponential change in the metabolic processes of embryonic elasmobranchs (Hoff 2008).

Development time of thorny skate (*Amblyraja radiata*), a species found at similar latitudes, could vary by as much as 1.5 years between populations differing in mean developmental temperatures (Berestovskii 1994). Developmental time is so sensitive to temperature in Alaska skate (*Bathyraja parmifera*) that an increase of 0.05°C in the mean environmental temperature can result in a 16% (~6 month) decrease in the developmental period (Hoff 2008). Although reduced developmental time in warmer conditions might improve survival probability to hatching, it can cause irregular coloration and patterning, skeletal abnormalities and increased metabolic rate and ventilation, which leads to a decline in overall fitness and significantly increases mortality rates of juvenile sharks (Rosa et al. 2014, Gervais et al. 2015). Early ontogenetic stages, incapable of thermotaxic behavior, will be most susceptible to climatic events (Pimentel et al. 2014).

Rising ocean temperatures associated with climate change are likely to have an effect on the recruitment and physiology of oviparous elasmobranchs and should be an important consideration for future conservation management plans in the North-East Atlantic. Elevated temperatures affecting the metabolic activity and foraging behavior of large marine predators, could have effects on ecosystem stability by altering the composition and distribution of important prey communities (Pistevos et al. 2016). According to the OSPAR commission (2009) and Morris et al. (2018), the North Sea surface temperature has warmed by 1-2°C over the last 25 years, during which time summer periods have become warmer and lasted longer, while winters have become shorter and milder. In the Irish Sea, temperatures are expected to increase by ~1.9°C over the 21<sup>st</sup> century, with shallow coastal water exhibiting the warmest temperatures, while deep channels remain cooler with less variability in temperature fluctuations and greater stratification between layers (Olbert et al. 2012). The rate of sea surface warming around the British Isles, excepting areas of stratification, has been reported to be up

to six times faster than the global average (Dye et al. 2013), with the region recognized as one of 20 global hotspots of marine climate change based on ocean temperature trends (Hobday & Pecl 2014).

An aspect of particular relevance to restoration of endangered species is that when the pejus temperature (the limit of optimal haemolymph oxygenation) of a species is exceeded the associated increased metabolic costs compromise growth, fitness, and so population increase (Neuheimer et al. 2011). This could differentially affect blue skate, which appear to occupy a more restricted temperature range. However, large, mature flapper skates have an apparent preference for cold, deep trenches, which could put them at higher risk of rising temperatures if these important habitats become too warm. Further, although skates are largely associated with the benthic environment, flapper skate are known to actively hunt pelagic prey and utilize the entire depth profile available (Wearmouth & Sims 2009, Neat et al. 2014,), which suggests changes in sea temperatures could impact its foraging behavior.

The thermal scopes observed in this study have implications with respect to projected climate change scenarios. Recent work has identified increases in maximum annual temperatures as drivers of recent population extinctions, unless compensated by species niche (Roman-Palacios & Wiens, 2020). If temperatures increase in the North-East Atlantic the blue skate's association with warmer waters may predict a likely expansion northward and to greater depths. The flapper skate's apparent tolerance of a wider range of temperatures suggests it could pursue a more flexibile strategy requiring less range shift, provided critical components of its ecosystem, such as prey species and nursery areas, are not adversely affected. However, if the temperatures

of their critical deep trench habitats increase with climate change, this could put the flapper skate at a greater disadvantage.

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### 6. LITERATURE CITED

Bendall VA, Hetherington SJ, Barreau T, Nicholson R, Winpenny L (2017) Common Skate
Survey Annual Report (ELECTRA MF6001: Work Package Task 1.4) Cefas. 30 pp
Berestovskii EG (1994) Reproductive biology of skates of the family Rajidae in the seas of the
far north. J Ichthyol 34:26-37
Biais G, Coupeau Y, Seret B, Calmettes B, Lopez R, Hetherington S, Righton D (2017)
Return migration patterns of porbeagle shark (*Lamna nasus*) in the Northeast Atlantic:
implications for stock range and structure. ICES Mar Sci 74: 1268-1276

Brander K (1981) Disappearance of the common skate from the Irish Sea. Nature 290:48-49

344	Di Santo V (2015) Ocean acidification exacerbates the impacts of global warming on
345	embryonic little skates, <i>Leucoraja erinacea</i> (Mitchill). J Exp Mar Biol Ecol 463:72-78
346	Doherty, PD, Baxter JM, Gell FR, Godley BJ, Graham RT, Hall G, Hall JHawkes LA,
347	Henderson SM, Johnson L, Speedie C, Witt MJ (2017a) Long-term satellite tracking
348	reveals variable seasonal migration strategies of basking sharks in the north-east Atlantic.
349	Sci Rep 7: 42837
350	Doherty, PD Baxter JM, Godley BJ, Graham RT, Hall G, Hall J, Hawkes LA, Henderson SM,
351	Johnson L, Speedie C, Witt MJ (2017b) Testing the boundaries: Seasonal residency and
352	inter-annual site fidelity of basking sharks in a proposed Marine Protected Area. Biol
353	Conserv 209: 68–75
354	Dulvy NK, Notobartolo di Sciara G, Serena F, Tinti F, Ungaro N, Mancusi C, Ellis J (2006)
355	Dipturus batis. In: IUCN 2011. IUCN Red List of Threatened Species. Version 2011.2
356	Dulvy N.K, Rogers S.I, Jennings S, Stelzenmüller V, Dye S.R, Skjoldal H.R (2008) Climate
357	change and deepening of the North Sea fish assemblage: a biotic indicator of warming
358	seas, J Appl Ecol 45, 1029-1039
359	Dye SR, Hughes SL, Tinker J, Berry DI, Holliday NP, Kent EC, Kennington K, Inall M, Smyth
360	T, Nolan G, Lyons K, Andres O, Beszczynska-Möller A (2013) Impacts of climate
361	change on temperature (air and sea). MCCIP Science Review: 1-12
362	Earl DA, vonHoldt BM (2012) STRUCTURE HARVESTER: a website and program for
363	visualizing STRUCTURE output and implementing the Evanno method. Conserv Genet
364	Resour 4:359-361
365	Ebert DA, Ebert TB (2005) Reproduction, diet and habitat use of leopard sharks, Triakis
366	semifasciata (Girard), in Humboldt Bay, California, USA. Mar Freshw Res 56:1089-
367	1098

368	Ellis JR, Cruz-Martinez A, Rackham BD, Rogers SI (2005) The distribution of chondrichthyan
369	fishes around the British Isles and implications for conservation. J Northwest Atl Fish
370	Sci 35:195-213
371	El Nagar A, McHugh M, Rapp T, Sims DW, Genner MJ (2010) Characterisation of
372	polymorphic microsatellite markers for skates (Elasmobranchii: Rajidae) from expressed
373	sequence tags. Conserv Genet 11:1203-1206
374	Estalles M, Coller NM, Perier MR, Di Giácomo EE (2011) Skates in the demersal trawl fishery
375	of San Matías Gulf: species composition, relative abundance and maturity stages. Aquat
376	Living Resour 24:193-199
377	Evans JS (2018) spatialEco. R package version 0.1.1-1. <a href="https://CRAN.R-">https://CRAN.R-</a>
378	project.org/package=spatialEco
379	Gervais C, Mourier J, Rummer JL (2015) Developing in warm water: irregular colouration and
380	patterns of a neonate elasmobranch. Mar Biodivers Rec 46:43-744
381	Griffiths A, Sims DW, Cotterell S, Nagar A, Ellis JR, Lunghammar A, McHugh M, Neat FC,
382	Pade N, Queiroz N, Serra-Pereira B, Rapp T, Wearmouth V, Genner MJ (2010)
383	Molecular markers reveal spatially segregated cryptic species in a critically endangered
384	fish, the common skate (Dipturus batis). Proc Roy Soc B 277: 1497-1503
385	Heesen HJL, Daan N, Ellis JR (2015) Fish Atlas of the Celtic Sea, North Sea and Baltic Sea:
386	Based on International Research Vessel Data. Wageningen Academic Publishers,
387	Wageningen, The Netherlands. 550 pp
388	Hobday AJ, Pecl, GT (2014) Identification of global marine hotspots: sentinels for change and
389	vanguards for adaptation ac- tion, Rev Fish Biol Fisher 24: 415-425

390	Hoff GR (2008) A nursery site of the Alaska skate (Bathyraja parmifera) in the eastern Bering						
391	Sea. Fish Bull 106:233-244						
392	Holt RE, Foggo A, Neat FC, Howell KL (2013) Distribution patterns and sexual segregation						
393	in chimaeras: implications for conservation and management. ICES J Mar Sci 70:1198-						
394	1205						
395	Hunter E, Buckley AA, Stewart C, Metcalfe JD (2005) Migratory behaviour of the thornback						
396	ray, Raja clavata, in the southern north sea J Mar Biol Assoc UK 85:1095-1105						
397	Iglesias SP, Toulhoat L, Sellos DY (2010) Taxonomic confusion and market mislabelling of						
398	threatened skates: important consequences for their conservation status. Aquatic						
399	Conserv: Mar Freshw Ecosyst 20:319-333						
400	IUCN (2011) IUCN 2008 Red List of Threatened Species. Available at:						
401	http://www.iucnredlist.org						
402	Jirik KE, Lowe CG (2012) An elasmobranch maternity ward: female round stingrays <i>Urobatis</i>						
403	halleri use warm, restored estuarine habitat during gestation. J Fish Biol 80:1227-1245						
404	Klimley AP (1985) Schooling in Sphyrna lewini, a Species with Low Risk of Predation: a Non-						
405	egalitarian State. Ethol 70:297-319						
406	Klimley PA (1987) The determinants of sexual segregation in the scalloped hammerhead shark,						
407	Sphyrna lewini. Environ Biol Fishes 18:27-40						
408	Kopelman NM, Mayzel J, Jakobsson M, Rosenberg NA, Mayrose I (2015) Clumpak: a program						
409	for identifying clustering modes and packaging population structure inferences across K.						
410	Mol Ecol Resour 15:1179-1191						

411	Last PR, Weigmann S, Yang L (2016) Changes to the Nomenclature of the Skates
412	(Chondrichthyes: Rajiformes). Rays of the World (Supplementary Information)
413	(November), 34 pp
414	Roman-Palacios C, Wiens, JJ (2020) Recent responses to climate change reveal the drivers of
415	species extinction and survival. PNAS 117(8): 4211-4217
416	McCutchen, M (2012) Population characterization of the common skate species complex in the
417	northeast Atlantic Ocean. MSc Thesis, University of Aberdeen, Aberdeen, UK
418	McKibben JR, Nelson DR (1986) Patterns of movements and grouping of grey reef sharks,
419	Carcharhinus amblyrhynchos at Enewetak, Marshal Islands. Bull Mar Sci 38:89–110
420	Morris, DJ, Pinnegar, JK, Maxwell, DL, Dye, SR, Fernand, LJ, Flatman, S, Williams, OJ
421	Rogers, SI (2018) Over 10 million seawater temperature records for the United Kingdom
422	Continental Shelf between 1880 and 2014 from 17 Cefas (United Kingdom government)
423	marine data systems. Earth Syst Sci Data 10: 27-51
424	Neat F, Pinto C, Buttett I, Cowie L, Travis J, Thorburn J, Gibb F, Wright PJ (2014) Site fidelity,
425	survival and conservation options for the threatened flapper skate (Dipturus intermedia).
426	Aquatic Conserv: Mar Freshw Ecosyst 25:6-20
427	Neuheimer AB, Thresher RE, Lyle JM, Semmens JM (2011) Tolerance limit for fish growth
428	exceeded by warming waters. Nature Clim. Change 1:110-113
429	Olbert AI, Dabrowski T, Nash S, Hartnett M (2012) Regional modelling of the 21st century
430	climate changes in the Irish Sea. Cont Shelf Res 41:48-60
431	Orlov AM, Tokranov AM (2010) Reanalysis of long-term surveys on the ecology and biology
432	of mud skate (Rhinoraja taranetzi Dolganov, 1985) in the northwestern Pacific (1993-
433	2002). J Appl Ichthyol 26:861-871

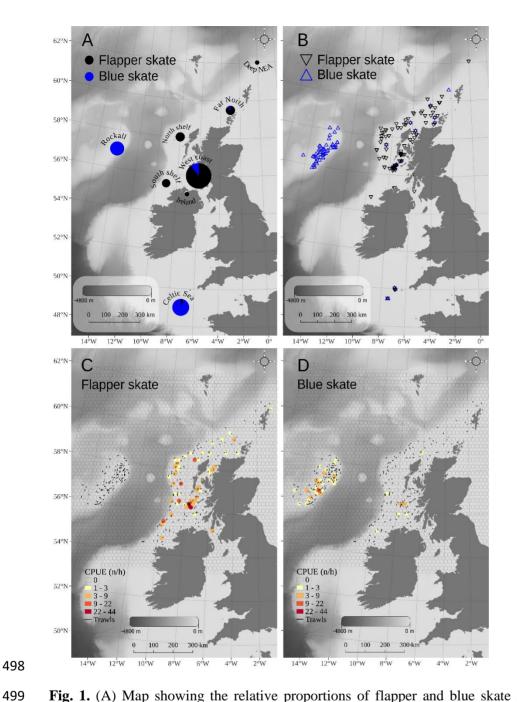
434	OSPAR Commission (2009) Impacts of climate change on the North-East Atlantic ecosystem.
435	Monitoring and Assessment Series
436	Pauly D, Christensen V, Guenette S, Pitcher TJ, Sumaila UR, Walters CJ, et al. (2002) Towards
437	sustainability in world fisheries. Nature 418:689-695
438	Perry AL, Low PJ, Ellis JR, Reynolds JD (2005) Climate change and distribution shifts in
439	marine fishes. Science 308(5730):1912-5
440	Pimentel MS, Faleiro F, Dionisio G, Repolho T, Pousao-Ferreira P, Machado J, Rosa R (2014)
441	Defective skeletogenesis and oversized otoliths in fish early stages in a changing ocean.
442	J Exp Biol 217:2062-2070
443	Pistevos JC, Nagelkerken I, Rossi T, Connell SD (2016) Antagonistic effects of ocean
444	acidification and warming on hunting sharks. Oikos 126:241-247
445	Pratt HL, Carrier JC (2001) A review of elasmobranch reproductive behavior with a case study
446	on the nurse shark, Ginglymostoma cirratum. Environ Biol Fishes 60:157-188
447	Pretorius C, Griffiths CL (2013) Patterns of egg deposition and egg development in the
448	catsharks Poroderma pantherinum and Haploblepharus pictus. Afr Zool 49:115-124
449	Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using
450	multilocus genotype data. Genetics 159:945-959
451	QGIS Development Team (2019) QGIS Geographic Information System. Open Source
452	Geospatial Foundation:http://qgis.org
453	Rodríguez-Cabello C, Sánchez F, Olaso I (2007) Distribution patterns and sexual segregations
454	of Scyliorhinus canicula (L.) in the Cantabrian Sea. J Fish Biol 70:1568-1586
455	R Core Team (2020) R: A Language and Environment for Statistical Computing. R Foundation
456	for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/

457	Rosa R, Baptista M, Lopes VM, Pegado MR, Paula JR, Trubenbach K, Leal MC, Calado R,
458	Repolho T (2014). Early-life exposure to climate change impairs tropical shark survival.
459	Proc Roy Soc B 281:1793
460	Royle JA, Chandler RB, Yackulic C, Nichols JD (2012) Likelihood analysis of species
461	occurrence probability from presence-only data for modelling species distributions.
462	Methods Ecol Evol 3:545–554
463	Sambrook J, Fritsch EF, Maniatis T (1989) Molecular cloning: A Laboratory Manual. 2nd
464	Edition. Cold Springs Harbor Press. Cold Springs Harbor
465	Saunders RA, Royer F, Clarke MW (2011) Winter migraton and diving behavior of the
466	porbeagle shark, Lamna nasus, in the Northeast Atlantic. ICES J Mar Sci 68:166-174
467	Schlaff AM, Heupel MR, Simpfendorfer CA (2014) Influence of environmental factors on
468	shark and ray movement, behavior and habitat use: a review. Rev Fish Biol Fish 24:1089-
469	1103
470	Sinclair BJ, Marshall KE, Sewell MA, Levesque DL, Willett CS, Slotsbo S, et al. (2016) Can
471	we predict ectotherm responses to climate change using thermal performance curves and
472	body temperatures?. Ecol Lett 19:1372-1385
473	Stevens JD, Bonfil R, Dulvy NK., Walker PA (2000) The effects of fishing on sharks, rays,
474	and chimaeras (chondrichthyans), and the implications for marine ecosystems. ICES J
475	Mar Sci 57:476-494
476	Tserpes G, Maravelias CD, Pantazi M, Peristeraki P (2013) Distribution of relatively rare
477	demersal elasmobranchs in the eastern Mediterranean. Estuar Coast Shelf Sci 117:48-53
478	Vaudo JJ, Heithaus MR (2013) Microhabitat selection by marine mesoconsumers in a
479	thermally heterogeneous habitat: behavioral thermoregulation or avoiding predation risk?
480	PLOS ONE 8: e61907

481	Walker PA, Hislop JRG (1998) Sensitive skate or resilient rays? Spatial and temporal shifts in
482	ray species composition in the central and north-western North Sea between 1930 and
483	the present day. ICES J Mar Sci 55:392-402
484	Wallman HL, Bennett WA (2006) Effects of parturition and feeding on thermal preference of
485	Atlantic stingray, Dasyatis sabina (Lesueur). Environ Biol Fishes 75:259-267
486	Wearmouth VJ, Sims DW (2009) Movement and behaviour patterns of the critically
487	endangered common skate Dipturus batis revealed by electronic tagging. J Exp Biol Ecol
488	380:77-87
489	Wickham H (2016) ggplot2: Elegant graphics for data analysis. Springer-Verlag New York

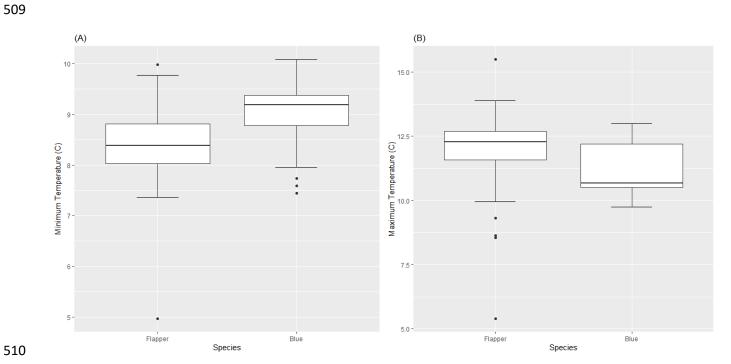
**Table 1.** Summary of the number of skate from each sampling area for flapper and blue skate after genetic confirmation. Bottom temperatures were averaged over a 12 month period for each sampling location, and average depth was calculated for areas where depth was recorded. Numbers outside brackets are total numbers used in distribution mapping, while numbers in brackets are those with bottom temperature measures for sampling locations. \* The Scottish west coast, including the Loch Sunart to Sound of Jura MPA. Data not available, na.

Sampling Area	Temp (C°)	Average Depth (m)	Dates	flapper skate	blue skate	Total
North Scotland (FN)	10.40	168.2	2011-2013	49	9	58
Deep NEA (DS)	5.11	na	2013	2 (1)	0	2 (1)
Ireland (IRE)	11.18	na	2013	7 (1)	0	7 (1)
West Coast* (SWC)	10.43	168.4	2011-2013	384 (324)	43	427 (367)
Rockall (ROCK)	10.65	200.5	2008, 2011- 2013	0	129	129
Shelf North (NSHLF)	10.21	185.8	2012-2012	56	0	56
Shelf South (SSHLF)	10.63	151.3	2012-2013	47	1	48
Celtic Sea (CS)	10.48	na	2011	9	179	188
Total	-		2008-2013	554 (487)	361	915 (848)



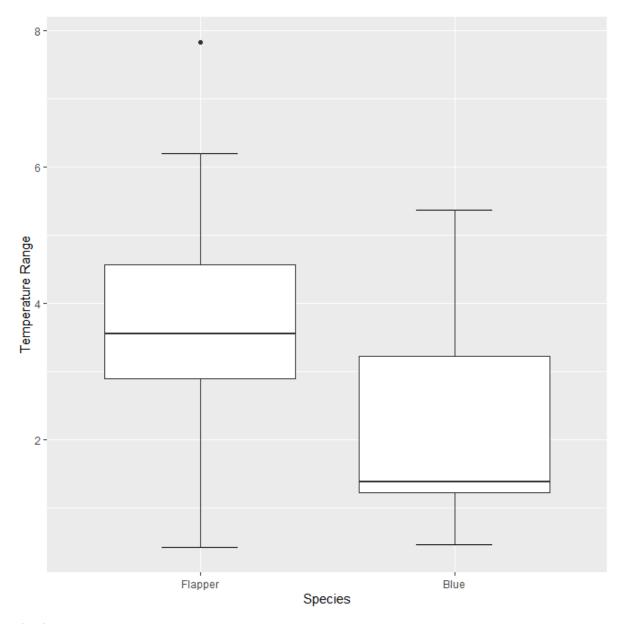
**Fig. 1.** (A) Map showing the relative proportions of flapper and blue skate sampled in this study. Sampling locations included: the Celtic Sea (n=188), the northern Scottish continental shelf (North Shelf; n=56), the southern Scottish and Irish continental shelf (South Shelf; n=48), the Rockall Bank (n=129), the western coast of Scotland (West Coast; n=427; includes the Loch Sunart to Sound of Jura MPA), Ireland (n=7), far north of Scotland in Orkney and Shetland (Far North; n=58) and a deep sea haul in the North-East Atlantic (Deep NEA; n=2). (B) Plot of capture locations of flapper and blue skate sampled in this study. (C) Map showing the catch per unit effort of flapper skate sampled in Marine Scotland Science surveys (D) Map showing the catch per unit effort of blue skate sampled in Marine Scotland Science surveys.





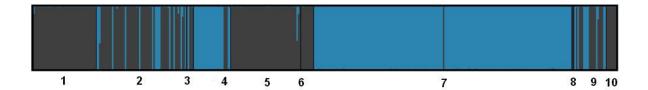
**Fig. 2.** Boxplots with error bars comparing the mean minimum (A) and maximum (B) temperatures ( $C^{\circ}$ ) characterizing areas where flapper skate and blue skate were found in this study. Box-plots include the median (solid line in box) and 25<sup>th</sup> and 75<sup>th</sup> percentiles; whiskers are the 10<sup>th</sup> and 90th percentiles with circles representing outliers.





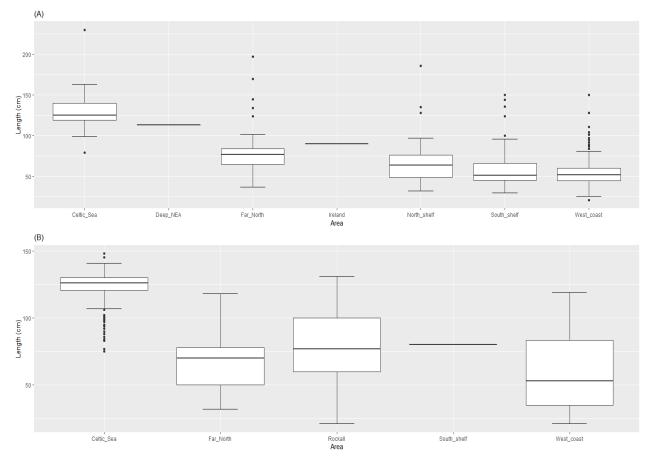
**Fig. 3.** Temperature ranges of areas where flapper skate and blue skate populations were found in this study. Boxplots with error bars show the mean range of temperatures for areas where each species was found. The box-plot includes the median (solid line in box) and 25<sup>th</sup> and 75<sup>th</sup> percentiles; whiskers are the 10<sup>th</sup> and 90th percentiles with circles representing outliers.

### 522 SUPPLEMENT



**Supplementary Fig. S1.** STRUCTURE outcome for K=2, indicating two clusters for species membership from 10 different surveys throughout the Northeast Atlantic. Each individual is represented by a vertical line; flapper skate are shown in gray and blue skate are blue. Numbers correspond to different research surveys in coastal waters of the British Isles. Research surveys included the following locations: north of Scotland (1, 3, 5, 8, 9), west coast or continental shelf of Scotland (2, 4, 5, 9) the Celtic Sea (6, 7) and Ireland (10). Two individuals, from survey 2 and 5, failed to amplify at all 5 microsatellites and were removed prior to any analyses (structure plot shows 404 samples).





**Supplementary Fig. S2.** Boxplots with error bars comparing the mean total length (in cm) of flapper skate (A) and blue skate (B) in sampling locations where each species was present. Locations include the Celtic Sea, Deep Northeast Atlantic (Deep\_NEA), Orkney and Shetland (Far\_north), Ireland, Rockall, Northern Scottish continental shelf (North\_shelf), Southern Scottish and Irish continental shelf (S\_shelf) and the West coast of Scotland (West\_coast). Box-plots includes the median (solid line in box) and 25<sup>th</sup> and 75<sup>th</sup> percentiles; whiskers are the 10<sup>th</sup> and 90th percentiles with circles representing outliers.

**Supplementary Table S1.** Summary of Mann-Whitney U tests performed between species and binomial tests to determine if sex ratios were equal. **Test** gives the variable that was tested, **n** is the sample size, **Stat** is the test statistic used, and the **P-value**, **confidence intervals** and **mean** for each species (where applicable) is also presented. Mann-Whitney U tests were run on the minimum (Min), maximum (Max) and temperature range (Range), while a 2-tailed binomial test was used to examine sex ratios in flapper skates and blue skates. NA is no data available.

Test	n	Stat	P-value	Confidence interval	Mean (flapper skate)	Mean (blue skate)
Min	173	W=5765	P<0.001	(0.51, 0.86)	8.41	9.07
Max	173	W=1704	P<0.001	(-1.52, - 0.89)	12.01	11.09
Range	173	W=1599	P<0.001	(-2.19, - 1.42)	3.59	2.01
Flapper sex ratio	487	Exact binomial test	P=0.415	(0.44-0.53)	NA	NA
Blue sex ratio	361	Exact binomial test	P=0.045	(0.50-0.61)	NA	NA