

1 **Calibration and validation of the DNDC model to estimate nitrous oxide emissions and**
2 **crop productivity for a summer maize-winter wheat double cropping system in Hebei,**
3 **China**

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15 Key words: Calibration; Validation; Nitrous oxide; DNDC model; Crop productivity; Summer
16 maize-winter wheat double cropping system.

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46

47 **Abstract**

48

49 The main aim of this paper was to calibrate and evaluate the DeNitrification-DeComposition
50 (DNDC) model for estimating N₂O emissions and crop productivity for a summer maize-winter
51 wheat double cropping system with different N fertilizer rates in Hebei, China. The model's
52 performance was assessed before and after calibration and model sensitivity was investigated.
53 The calibrated and validated DNDC performed effectively in estimating cumulative N₂O
54 emissions (coefficient of determination (1:1 relationship; r^2) = 0.91; relative deviation (RD) =
55 -13 to 16%) and grain yields for both crops (r^2 = 0.91; RD = -21 to 7%) from all fertilized
56 treatments, but poorly estimated daily N₂O patterns. Observed and simulated results showed
57 that optimal N fertilizer treatment decreased cumulative N₂O flux, compared to conventional
58 N fertilizer, without a significant impact on grain yields of the summer maize-winter wheat
59 double cropping system. The high sensitivity of the DNDC model to rainfall, soil organic
60 carbon and temperature resulted in significant overestimation of N₂O peaks during the warm
61 wet season. The model also satisfactorily estimated daily patterns/ average soil temperature (°
62 C; 0-5 cm depth) (r^2 = 0.88 to 0.89; root mean square error (RMSE) = 4° C; normalized RMSE
63 (nRMSE) = 25% and index of agreement (d) = 0.89-0.97) but under-predicted water filled pore
64 space (WFPS; %; 0-20 cm depth) (r^2 = 0.3 to 0.4) and soil ammonium and nitrate (exchangeable
65 NH₄⁺ & NO₃⁻; kg N ha⁻¹; r^2 = 0.97). With reference to the control treatment (no N fertilizer),
66 DNDC was weak in simulating both N₂O emissions and crop productivity. To be further
67 improved for use under pedo-climatic conditions of the summer maize-winter wheat double
68 cropping system we suggest future studies to identify and resolve the existing problems with
69 the DNDC, especially with the control treatment.

70

71 Capsule

72 The calibrated DNDC model effectively estimated cumulative N₂O emissions, grain yields
73 and soil temperature but underestimated WFPS and soil N, in a winter wheat-summer maize
74 double cropping system.

75 Key words: Calibration; Validation; Nitrous oxide; DNDC model; Crop productivity; Summer
76 maize-winter wheat double cropping system.

77

78 **1 Introduction**

79

80 Quantification of greenhouse gas (GHG; CO₂, CH₄ and N₂O) emissions from agricultural soils
81 is essential for developing mitigation options and policies. However, this requires establishing
82 and maintaining field flux measurement sites which are time consuming and expensive. Well-
83 calibrated simulation models for GHG emissions offer an opportunity to complement physical
84 experiments by employing computers to calculate the likely outcomes of different physical
85 phenomenon (Giltrap et al., 2010). Nitrification and denitrification are the main processes
86 responsible for N₂O production in soils and their contribution depends on the environmental
87 conditions (Mathieu et al., 2006). Simulation models have the ability to simulate relationships
88 between soil physical, chemical and microbial processes that underpin nitrification,
89 denitrification and decomposition. They also allow complex interactions and real-world
90 problems to be examined in a time effective way, by applying mathematical knowledge and
91 computational power. Moreover, simulation models can support decision makers by
92 facilitating the understanding of a system and allow potential mitigation strategies of GHG
93 emissions, and a range of climate change-land use change scenarios to be examined (Giltrap et
94 al., 2010).

95 Simulation models are very diverse and range from simple empirical relationships
96 based on statistical analyses to complex mechanistic models that consider numerous soil-
97 climate-crop parameters controlling and influencing GHG production and emissions from soils
98 (Roelandt et al., 2005; Jinguo et al., 2006). The exact estimation of the trace GHG, nitrous
99 oxide (N₂O), emissions from soil is difficult and represents a challenge for most of the models
100 which perform over a wide range of conditions. However, soil parameters and almost all
101 processes responsible for production, consumption and transport of this gas can be simulated
102 (Williams et al., 1992). One of the process models used to estimate N₂O emissions is the
103 DeNitrification-DeComposition (DNDC) model. The DNDC model is a biogeochemical model

104 used to estimate soil GHG emissions and crop production. Although it was initially developed
105 for conditions in the USA (Li et al., 1992, 2000), it has been used for simulating N₂O emissions
106 worldwide e.g. in Canada (Smith et al., 2010), Europe (Kesik et al., 2006; Abdalla et al., 2009)
107 and extensively in China (Deng et al., 2011; Hu et al., 2012).

108 China is facing the dual challenge of increasing crop production for its growing
109 population while at the same time reducing its GHG emissions. Therefore, a plan for improving
110 agricultural management practices to promote grain yields and minimize GHG emissions is
111 needed (Chen et al., 2014). Two of the primary cereal crops in China are maize and wheat
112 which are grown on an area of about 42 and 24 million ha (FAO, 2017), respectively. Maize
113 is also an important forage crop, where about 68% of its production in China is used for animal
114 feed (Ely et al., 2016). Summer maize-winter wheat double cropping system is a common
115 cropping system in the North China plain. Previous studies found that crop rotation/ double
116 cropping system positively increased crop yields compared to monoculture management (Laik
117 et al., 2014). However, both the maize and wheat crops require a large amount of N fertilizer
118 for optimum growth and production. In addition, farmers commonly overuse N fertilizer or
119 apply a low efficiency types (Li et al., 2012). They usually add 30-60% more N fertilizers than
120 the level required for optimum crop yields (Norse, 2011). However, overuse of N fertilizer has
121 recently started to decline in some areas and the government set a policy of zero growth in N
122 fertilizer and pesticide use by 2020 (Powlson et al., 2018).

123 Nitrous oxide is a potent GHG. The emission of this gas from agriculture is produced
124 through biological processes in soils and the degree of variation (spatial and temporal) in the
125 emissions depends on soil type, land use and climatic factors (e.g. rainfall, temperature)
126 (Conrad, 1996). The inorganic N pool provides electrons for producing energy during
127 nitrification whilst, organic C provides electrons to reduce combined N during denitrification
128 (Addiscott et al., 1983; Khalil et al., 2002). Unfavourable management practices result in high
129 N₂O emissions which are mainly controlled by available N and C in soils (Galloway 1998;
130 Ding et al. 2007). Management can also influence soil fertility, indirectly, through
131 management-induced changes in plant composition (Collins et al., 1998; Patra et al., 2006) and
132 thereby, increase gas fluxes.

133 Modelling of a double / multiple cropping system is still a challenge because of the
134 hysteresis influence on soil properties such as soil moisture, nutrients and soil organic C (SOC).
135 Over the past 25 years many developments have been made to the DNDC model to meet the
136 needs of users. These include, among others, modularization of the code structure (Haas et al.
137 2013), and development of an integral optimisation function for crop and other input

138 parameters (Lamers et al., 2007; Van Oijen et al., 2011). However, to the best of our
139 knowledge, the model has not previously been calibrated for a summer maize-winter wheat
140 double cropping system in China. The main aim of this paper was to calibrate and evaluate the
141 DNDC model for estimating N₂O emissions and crop productivity for a summer maize- winter
142 wheat double cropping system with different N fertilizer rates in Hebei province, the North
143 China plain. Additionally, the ability of the model to estimate soil variables of temperature,
144 water filled pore space (WFPS) and soil N (exchangeable NH₄⁺ and NO₃⁻) was assessed.
145 Results are discussed in terms of highlighting the strengths, weaknesses and potential future
146 improvements to the DNDC model for simulating the double cropping system in China.

147

148 **2 Materials and methods**

149

150 **2.1 Experimental site**

151 This study used the data published in Song et al. (2018) to calibrate and validate the DNDC
152 model. An experiment was set up in Quzhou county, Hebei province, to investigate the impacts
153 of N management on N₂O emissions. As detailed in Table S1, five N treatments with four
154 replicates in a fully randomized block design were investigated. These treatments were: control
155 (no N fertilizer); conventional N (the amount of N fertilizer used in current practice; see Table
156 S1); the other three treatments were designed with optimized fertilizer N rates, namely: optimal
157 N; 0.7*optimal N and 1.3*optimal N fertilizer (*= means multiplication). Optimal N fertilizer
158 was calculated by the in-season root zone N management strategy to mitigate GHG emissions
159 (Cui et al., 2013). Here, soil N (NH₄⁺-N and NO₃⁻-N) in the root zone was subtracted from the
160 target N values for the growing period. Further details about the site, crop, soil parameters and
161 management are shown in Song et al. (2018).

162

163

164 **2.2 Field measurements**

165 **2.2.1 Temperature and precipitation**

166 Mean daily air temperature and precipitation were collected from the weather station at the
167 study site (Fig. S1) as described by Song et al. (2018).

168

169 **2.2.2 Fluxes of N₂O**

170 Measurements of N₂O fluxes were carried out throughout the experimental period from June
171 2012 to June 2014, using the closed static chamber method. Gas samples were collected on a

172 daily basis for 10 days after application of N fertilizer and 3 days after irrigation or rainfall
173 (>20 mm). However, for the remaining periods, the gas was sampled every 4 days, except in
174 winter when the gas was sampled weekly. More details about N₂O measurements can be found
175 in Song et al. (2018).

176

177 **2.2.3 Calculation of N₂O flux**

178 The daily N₂O flux was calculated as shown in Song et al. (2018).

179

180

181 **2.2.4 WFPS (%) and soil N (exchangeable NH₄⁺ and NO₃⁻)**

182 Soil samples for measurements of WFPS and mineral N (exchangeable NH₄⁺ and NO₃⁻) were
183 collected and calculated as described in Song et al. (2018).

184

185

186 **2.3 Model description**

187 DNDC v. 9.5 is a biogeochemistry model which describes the soil C and N cycles and GHG
188 fluxes from agricultural systems (Gilhespy, 2014). The DNDC model accommodates six sub-
189 models (Li et al., 1992, 2000).

190

191 **2.4 Model's calibration and sensitivity analysis**

192 This study represents a further step of our previous studies to investigate the suitability of the
193 DNDC model for estimating N₂O, crop yield and soil properties for China's cropland (Song et
194 al., 2018; Yue et al., 2018). The DNDC model was calibrated to produce measured crop yields
195 / cumulative N₂O emissions for the site using the measured data from the 0.7 * optimal N
196 treatment. Data from the control plot were not used for calibration because there were many
197 days in the control data in which the measured N₂O flux was negative and negative fluxes are
198 not simulated by DNDC.

199 Model calibration for crop yields and cumulative N₂O emissions was done by
200 optimizing a combination of different crop growth parameters (maximum biomass production,
201 biomass fraction, biomass C/N ratio, thermal degree days, water demand and optimum
202 temperature) and adjusting SOC inputs, respectively. Different crop parameters/ SOC input
203 default values were tested until the model matched the measured grain yield/ cumulative N₂O
204 flux values (Table 1). The grain yield was measured in t ha⁻¹. The calibrated model was then
205 used to run the other 4 treatments (control, conventional N, optimal N and 1.3 * optimal N).

206 The sensitivity of the DNDC model and the attribution of N₂O and summer maize/ winter
207 wheat grain yields to different input parameters were investigated to quantify the effects of
208 these parameters on the N₂O emissions and grain yields (Smith and Smith, 2007; Abdalla et
209 al., 2009a). We change only one parameter at a time and kept the other ones constant.
210 Simulations were run to assess how N₂O and grain yields were influenced by different climate
211 parameters: average daily temperature (increased/ decreased by a range from 1 to 3° C with an
212 increment of 1° C) and average daily rainfall (increased/decreased by a range from -30% to
213 +30% with an increment of 10%). The model was also run to see how N₂O and grain yields
214 were affected by changes in SOC and for the amount of N fertilization rate and water
215 irrigation. SOC, N fertilizer and irrigation were changed by -30% to +30% with an increment
216 of 10%.

217

218 **2.5 Model run, validation and statistical evaluation**

219 To run the DNDC model, climate, soil and management data including N fertilizer, irrigation
220 and tillage were input into the model. These are summarized in Tables 1, 2 and 3. The model
221 testing was carried out by comparing (1) simulated and observed daily/ cumulative N₂O fluxes
222 (2) simulated and observed crop grain yields and (3) simulated and observed soil N
223 (exchangeable NH₄⁺ and NO₃⁻) (4) simulated and observed soil moisture in terms of WFPS (5)
224 simulated and observed soil temperature. The model was validated by comparing observed and
225 simulated values.

226 The model accuracies were evaluated by calculating root mean square error (RMSE;
227 equation 1), normalized RMSE (nRMSE; equation 2), index of agreement (d; equation 3) Yang
228 et al. 2014) and modelling efficiency (EF; equation 4) (Nash and Sutcliffe, 1970). Using these
229 indices help us to quantify the overall model performance. The RMSE have the same unit of
230 simulated and observed values, whilst nRMSE is a relative measure. The d (0 ≤ d ≤ 1) gives
231 the degree of deviation towards zero. EF (- ∞ to 1) compares the ability of the model to
232 reproduce the daily data variability based on the arithmetic mean of the measurements.
233 Negative EF value shows a poor performance, a value of 0 indicates that the model does not
234 perform better than using the mean of the observations, and values close to 1 indicate a ‘near-
235 perfect’ fit.

236

$$237 \quad RMSE = \sqrt{\frac{\sum_{i=1}^n (S_i - M_i)^2}{n}} \quad (1)$$

238

$$nRMSE = \frac{RMSE}{\bar{M}} \times 100 \quad (2)$$

239

240

$$d = 1 - \frac{\sum_{i=1}^n (S_i - M_i)^2}{\sum_{i=1}^n (|S_i - \bar{M}| + |M_i - \bar{M}|)^2} \quad (3)$$

241

242

$$EF = 1 - \frac{\sum_{i=1}^n (S_i - M_i)^2}{\sum_{i=1}^n (M_i - \bar{M})^2} \quad (4)$$

243

244

245 The relative deviation (RD; %) of the observed values from modelled ones was also calculated
246 as follow:

247

$$RD = (M_i - S_i) / M_i \quad (5)$$

248

249
250 Where S_i is the simulated value, M_i is the measured value, n is the number of measured values,
251 and \bar{M} is the average of the measured values. Cumulative flux for models results were
252 determined by the summation of modelled daily emissions over the experimental period (Cai
253 et al., 2003). Additionally, coefficient of determination (r^2), which is the correlation between
254 simulated and observed values was used to assess whether simulated values follow the same
255 pattern as observed values.

256

257 **3 Results**

258 **3.1 Model's calibration**

259 The adopted combination of crop parameters used for DNDC- calibration was shown in Table
260 2. The calibrated DNDC model successfully produced the exact measured crop yields ($t\ ha^{-1}$)
261 of the 0.7*optimal N treatment for each crop/ season. Likewise, the input amount of SOC at
262 0-10 cm in the model was adjusted to $0.021\ kg\ C\ kg^{-1}\ soil$ (i.e. SOC value resulted from the
263 model calibration) and the model also gave the measured cumulative N_2O flux for the 0.7*
264 optimal N treatment of $5.4\ kg\ N_2O-N\ ha^{-1}$.

265

266 **3.2 Model sensitivity analysis**

267 The sensitivity of the DNDC-model to the essential input parameters (i.e. rainfall, air
268 temperature, SOC, N fertilizer rate and water irrigation) for simulating cumulative N₂O flux
269 for the summer maize-winter wheat double cropping system was tested. The model was found
270 to be sensitive to changes in all of these parameters but to different extents (Fig. 1). The greater
271 response was to rainfall, where changing daily rainfall by a range from -30% to 30% changed
272 the cumulative N₂O emissions by a range from -50% to 42%. Changing SOC by a range from
273 about -30% to 30% changed cumulative N₂O emissions by a range from -36% to 39%. The
274 DNDC was also sensitive to changes in daily air temperature (°C) and N fertilizer application
275 rate. Changing daily air temperature and N fertilizer by a range from -3 °C to 3°C and from -
276 30% to 30% changed cumulative N₂O by ranges of -16% to 12% and -22% to 12%,
277 respectively. However, the model was less sensitive to irrigation where changing irrigation by
278 a range from -30% to 30% changed cumulative N₂O emissions by a range from -1% to 2%,
279 respectively. Here, increasing water irrigation had slight negative influence on the cumulative
280 N₂O emissions from soil.

281

282 **3.3 Evaluation of the DNDC model**

283 **3.3.1 Nitrous oxide emissions**

284 The DNDC model was able to predict timing of the daily observed N₂O flux peaks from all N
285 treatments during the two crop rotations, with few exceptions, but significantly overestimated
286 their magnitude (Fig. 2). These peaks appeared for all treatments including the controls on
287 occasions where combinations of higher daily rainfall (mm) and air temperature (°C) were
288 observed. For the control treatment, observed and simulated N₂O flux peaks corresponded to
289 higher daily rainfall and air temperature. However, the height of these peaks increased further
290 relative to the amount of the N fertilizer added in each N treatment plot. The highest observed
291 and simulated peaks were 6, 819, 149, 246 g N₂O-N ha⁻¹ d⁻¹ and 267, 831, 670 and 714 g N₂O-
292 N ha⁻¹ d⁻¹ for the control, conventional N, optimal N and 1.3 *optimal N, respectively. For all
293 treatments, RMSE ranged from 0.55 to 2.59 g N₂O-N ha⁻¹ d⁻¹; nRMSE from 4 to 20%, d from
294 0.10 to 0.50 and EF was <0 (Table 2). Both the observed and simulated cumulative N₂O flux
295 showed lower emissions from the optimal N fertilizer treatment compared to the conventional
296 and 1.3*optimal N fertilizer treatments (Table 2). The model performed better, for both N
297 fertilized and control treatments, after calibration compared to before calibration. Here, RD
298 ranged from -13 to 16% compared to -46 to -54% for the N fertilized treatments, respectively
299 (Table 2). However the model, generally, simulated daily/ cumulative N₂O flux for the control
300 in both cases, poorly. The DNDC overestimated the flux for the control treatment by 68%

301 before model calibration and by 42% after calibration. Overall, the model simulated cumulative
302 annual N₂O emissions from the maize-wheat double cropping system with an r² of 0.91 (1:1
303 relationship; Fig. S2).

304

305 **3.3.2 Crop yields**

306 With the exception of the control treatment, the DNDC model estimated observed grain yield
307 from both crops (summer maize and winter wheat) and all N treatments, effectively. The model
308 performed better after calibration, for both crops, compared to before calibration. For the N
309 treatments, the RD for simulating summer maize and winter wheat after calibration ranged
310 from -7 to 7% and from -21 to 6% compared to from 5 to 20% and from -42 to 59% before
311 calibration, respectively. The RD for simulating summer maize and winter wheat for the control
312 treatment after calibration ranged from -30% to -40% for the summer maize and from -50 to -
313 60% for the winter wheat compared to -92% to -97% and -83% to -87% before calibration,
314 respectively (Table 3). A 1:1 relationship showed that the DNDC simulated grain yield for
315 summer maize with r² of 0.89 and r² of 0.92 for winter wheat. The overall r² of simulated and
316 observed grain yields was 0.91 (Table 3; Fig. S3). On average, both the observed and simulated
317 grain yields showed that the optimal N fertilizer treatment slightly reduced crop yields (by 1 to
318 2%) compared to the conventional and 1.3* optimal fertilizer treatments (Table 3).

319

320 **3.3.3 Soil properties**

321 The daily WFPS (%) during the experimental period was primarily driven by rainfall. Both the
322 observed and simulated daily WFPS (%) corresponded well with increasing and decreasing of
323 daily rainfall. The DNDC model simulated daily trends in WFPS (%; 0-20 cm depth) with
324 some under-estimations of the observed values. 1:1 relationships showed that the model
325 simulated fluctuations in WFPS% (0-20 cm depth) with r² ranging from 0.3 to 0.4 (Fig. S4).
326 For all treatments the RD ranged from -62 to -76%. RMSE ranged from 12.9 to 42% and
327 nRMSE from 24 to 74. The d values were ranged from 0.40 to 0.75 and EF from <0 to 0.10.

328 With exception of the control treatment, the DNDC model was able to estimate timing
329 of soil N (exchangeable NH₄⁺ and NO₃⁻) peaks throughout the two rotations and all N
330 treatments, reasonably well, although it poorly estimated their magnitude (Fig. 3). The model
331 under-estimated the observed soil N peaks during periods of N application. The r² between the
332 daily observed and simulated values ranged from 0.11 to 0.17 and was 0.97 for the cumulative
333 soil N (1:1 relationship; Fig. S5). The RD ranged from -19 to -42% and RMSE ranged from
334 0.27 to 2.39 kg N ha⁻¹. The nRMSE values were small (2-4%); and d values were large (0.57-

335 0.75). The model significantly underestimated soil N for the control: (RD = -0.91; RMSE=
336 0.54 kg N ha⁻¹; nRMSE= 4% and d= 0.58 and EF ranged from <0 to 0.58 (Table 3; Fig. 3).

337 The DNDC model simulated daily trends in soil temperature (0-5 cm depth) throughout
338 the two summer maize-winter wheat double cropping system, effectively with some slight over/
339 under-estimation of the observed values (Fig. 4). The variation in measured soil temperature,
340 over the experimental period, was primarily derived by air temperature at the site. Both the
341 observed and simulated soil temperatures at 0-5 cm depth were not significantly different
342 between the different N treatments. The model simulated fluctuations in temperature (0-5 cm)
343 during the wet season (i.e. summer months) better than during the dry season (i.e. winter
344 months) (Figs. 1 and 5). A 1:1 relationship showed that the r² between the simulated and
345 observed values ranged from 0.88 to 0.89 (Fig. S6) and overall RD was 20%. The EF ranged
346 from 0.79 to 0.96 and RMSE was 4.1° C and both nRMSE and d values were reasonable; 25%
347 and 89-97, respectively (Table 3).

348

349 **4 Discussion**

350

351 **4.1 Model calibration and sensitivity analysis**

352 In this study, calibration and validation of the DNDC model using 0.7*optimal N treatment
353 was required because of the differences in the crop types and environment (i.e. DNDC was
354 originally developed for crop growth and environment in the USA). The calibration of DNDC,
355 especially for crop growth, is critically important due to the greater impacts of cropping
356 systems on soil N, C and water dynamics and thereby on the daily/ cumulative values of N₂O
357 emissions and other biogeochemical processes (Zhang and Niu 2016). The use of the
358 0.7*optimal N treatment, for which there are independent data, for model calibration was
359 essential. Many previous studies recommended calibration and validation of the DNDC model
360 to improve the accuracy of the model key biogeochemical processes (e.g. Tonitto et al. 2007;
361 Li et al. 2014). Our calibrated and validated model gave better estimation for cumulative N₂O
362 flux and crop grain yields.

363 The model sensitivity analysis for simulating N₂O flux showed that the DNDC model
364 is very sensitive to some climate, soil and management parameters including rainfall,
365 temperature, N fertilizer and SOC but less sensitive to water irrigation rate as shown in Fig. 1.
366 The DNDC was more sensitive to these parameters than in the study reported by Abdalla et al.
367 (2009a). This may be due to differences in the DNDC versions applied, soil texture,
368 management and environmental variables of the two sites. Rainfall increases both field

369 measured/ simulated soil moisture and thereby stimulates soil denitrification by lowering
370 oxygen dispersal into the soils (Abdalla et al. 2009b; Song et al. 2019). It also makes soil
371 organic C and nitrate more prone to denitrification processes by increasing their solubility
372 (Bowden and Bormann 1986). Therefore, rainfall events result in higher N₂O flux peaks/
373 cumulative flux as shown by Ludwig et al. (2011), Abdalla et al. (2012) and others. Water
374 irrigation also stimulates N₂O emissions (Yan et al. 2015). However, increasing water irrigation
375 rate can result in conditions of a complete denitrification in which N₂O is further reduced to N₂
376 (Conrad 1994) and consequently decrease N₂O emissions. This is why slightly negative effects
377 on the N₂O flux were observed in this study. In a two year study Kuang et al. (2018) reported
378 that flood irrigation decreased N₂O emissions, compared to drip irrigation, in one year and had
379 no significant difference in the second year.

380 Similar DNDC sensitivity to the higher air temperature found in this study, was also
381 reported by Abdalla et al. (2009a). This is interesting, and could result in significantly higher
382 N₂O emissions in the future especially because North China (area of this study) is projected to
383 change towards warmer and more humid conditions, and both rainfall and temperature will
384 increase as reported by Chu et al. (2017). The DNDC was sensitive to both additional synthetic
385 N fertilizer input and SOC. Changes in the amount of N fertilizer application rate has a direct
386 and a strong impact on N₂O emissions by making N available for the processes of nitrification
387 and denitrification in soils (Baggs and Blum, 2004). The N released to the atmosphere rely on
388 the amount of N used up by the crop (Abdalla et al., 2010). However, the overuse of N fertilizer
389 and application of a low use efficiency types in China (Li et al., 2012), if it continues, would
390 worsen the situation further. We found that the optimal N fertilizer treatment decreased
391 cumulative N₂O flux, compared to conventional and 1.3*optimal N fertilizer treatments,
392 without having a significant impact on grain yields of either crop. Hu et al. (2012) reported that
393 splitting the fertilizer into more applications reduced N₂O emissions from spring maize.
394 Moreover, using the same data used in this study, Song et al. (2018) found that cumulative and
395 yield-scaled N₂O emissions increased exponentially as N applications were raised above the
396 optimum rate in maize (*Zea mays* L.) and have quadratic increases in winter wheat (*Triticum*
397 *aestivum* L.).

398

399 **4.2 Evaluation of the DNDC model for simulating crop rotation**

400 **4.2.1 Nitrous oxide emissions**

401 In this study, although the DNDC correctly simulated the timing of most daily N₂O flux peaks
402 from all N treatments, it significantly overestimated their magnitudes. These peaks appeared

403 also in the control treatment and corresponded to combinations of higher daily rainfall and
404 temperature (the model is very sensitive to both parameters). Similar peaks at higher daily
405 rainfall events and temperature were simulated by Ludwig et al. (2011) and Abdalla et al.
406 (2012). These factors stimulate N₂O fluxes as they provide more substrate and favourable
407 conditions for both denitrification and nitrification in soils (Abdalla et al., 2014). Davidson et
408 al. (1993) and Huang et al. (2014) reported that under dry climate and low soil moisture,
409 nitrification was the main process behind N₂O production. The magnitude of the flux peaks
410 increased relative to the amount of added N in each treatment with the largest peak appearing
411 in the conventional N, and the lowest peak in the optimal N treatment. Li et al. (2012) reported
412 that avoiding application of N fertilizers coincident with heavy rainfall events can reduce N₂O
413 emissions from spring maize production in Northeast China. However, to reduce measured/
414 simulated N₂O emissions without significantly affecting crop yield, application of N fertilizer
415 should be decided depending on N available in soil and that removed by the crop (Wagner-
416 Riddle et al., 2007). The addition of N fertilizer stimulates nitrification and denitrification
417 processes and thereby, increases both observed and simulated N₂O emissions (Abdalla et al.,
418 2010; Abdalla et al., 2012). The significant differences between the simulated and observed
419 daily N₂O fluxes peaks resulted in a somewhat poor correlation between the daily simulated
420 and observed values. Generally, the field/ simulated N₂O peak emission events can account for
421 approximately 50-90% of the yearly emissions (Parkin and Kaspar, 2006; Wolf et al., 2010;
422 Abdalla et al., 2014). However, both the observed and simulated values do provide some
423 insight into likely peaks and trends in N₂O flux under different N management regimes. The
424 model imperfectly estimated the cumulative flux for the control treatment (RD = 42%) as a
425 result of poor estimation of WFPS (%), soil nitrate and crop yield under the control. One of the
426 disadvantages of the DNDC is that the model does not simulate negative N₂O flux values as in
427 the observed flux and therefore, overestimated the simulated flux. Another disadvantage is that,
428 the model under-estimated the observed WFPS (%) which is an important determinant of N₂O
429 flux (Dobbie and Smith, 2001). The WFPS (%) is one of the key requirements for a reliable
430 simulation of N₂O (Frolking et al., 1998), as changing its value may reduce the contribution of
431 simulated nitrification/ denitrification processes (Li et al., 2001). Moreover, the high sensitivity
432 of the DNDC model to rainfall events, SOC and temperature rendered the model less accurate
433 since it simulated many higher N₂O peaks that were not observed in the field. Uncertainties in
434 the observed values were also possible due to the limited number of field measurements
435 (Parkin, 2008) as N₂O is released in pulses from soils to the atmosphere (Hastings et al., 2010)
436 and peaks may appear for a maximum of few weeks only (Bell et al., 2012). Khalil et al. (2016)

437 reported that it is important to use a robust measurement protocol to get accurate validation of
438 the DNDC model in response to different management practices.

439 In this study, the DNDC model generally overestimated the cumulative observed N₂O
440 flux from the N treatments by an overall average of 13%. However, as the seasonal/ annual
441 cumulative N₂O fluxes were calculated by the interpolation method, and due to the fact that the
442 N₂O gas is characterized by episodic emissions, the observed cumulative emission could have
443 high uncertainties. Ju et al. (2011) reported that a sampling frequency of 3 or 6 days resulted
444 in an overestimation ranged from 112 to 228% in the total flux. According to Zhang et al.
445 (2002), the present version of DNDC is qualified for incorporating crop residue in the soil and
446 at the end of growing seasons. Residue turnover influences amounts of C and N added to the
447 soil and thereby, N₂O emissions. Previous studies have also shown an increase in simulated
448 N₂O flux due to the incorporation of cover crop residues into soils (Aulakh et al., 1984; Xiong
449 et al., 2002; Sarkodie-Addo et al., 2003). They justified that by the extra energy available for
450 denitrification, although provision of soil N through mineralisation of crop residues must also
451 be considered.

452

453 **4.2.2 Crop yields**

454 The DNDC model estimated crop grain yield for all N treatments effectively. However, the
455 model had difficulties in correctly estimating crop yield for the control treatment. This was due
456 to significantly under-predicting of both soil nitrate and WFPS (%) for the control treatment.
457 Additionally, the inability of the DNDC to correctly simulate the plant growth, although
458 improved by calibration, was a potential source of yield reductions in the control treatment (Hu
459 et al., 2017). Moreover, Abdalla et al. (2014) suggested improving the simulation of crop yield
460 by developing the crop growth module to include degree days of phenology stages and
461 radiation use efficiency for defining the growth curves for the crop. A new algorithm to the
462 crop sub-model was introduced by Zhang et al. (2002) for the China-DNDC-online, and acts
463 as an alternative approach to the empirical crop growth sub-model employed in DNDC (Li et
464 al. 1994). Reasonable simulation of crop yield is of key importance to accurately predict N₂O
465 emissions for process-based models of plant-soil systems.

466

467 **4.2.3 Soil properties**

468 The DNDC model effectively simulated soil temperature (0-5 cm depth) from the summer
469 maize-winter wheat double cropping system with r^2 ranging from 0.96 to 0.97. This is
470 comparable with the previously published studies of DNDC-temperature simulations under

471 crop multiple cropping system carried by Cui et al. (2014), Uzoma et al. (2015) and Li et al.
472 (2017). Cui et al. (2014) found r^2 ranged from 0.97 to 1.0, whilst Li et al. (2017) reported r^2
473 ranged from 0.89 to 0.97 between simulated and observed soil temperature for 0-5 cm and 0-
474 10cm depth, respectively. The model successfully predicted observed soil temperature by
475 tracing heat transfer between the different soil layers driven by soil heat capacity, temperature
476 gradient and heat conductivity. Our study revealed that the present algorithm in DNDC is
477 capable of correctly simulating soil temperature for double cropping system. This is important
478 because the ability of the model to simulate soil temperature is essential for simulating GHG
479 emissions, especially N_2O emissions. Soil temperature influences decomposition of soil
480 organic matter and response of soil microorganisms to other perturbations, such as the amount
481 of N fertilization and rainfall at the site (Wennman and Katterer, 2006). Likewise, accumulated
482 soil temperature is the main driver behind plant growth in the DNDC model. Plant growth
483 directly governs C and N contents and water in soils and, therefore, it is crucial to be simulated
484 correctly (Hu et al., 2012).

485 The DNDC model simulated WFPS (%) for all N treatments satisfactorily but was less
486 effective than that for simulating soil temperature (0-5 cm depth). The model under-estimated
487 the WFPS (%) and this increased the uncertainties associated with N_2O simulations and
488 resulted in poor fit with the observed flux (Wattenbach et al., 2010). The WFPS (%) determines
489 if a soil is anaerobic or aerobic by influencing the concentration and transport of oxygen
490 through the soil matrix (Song et al., 2019). Anaerobic conditions stimulate denitrification and
491 result in much higher production rates of N_2O (Ussiri and Lal, 2012). In contrast, Kuang et al.
492 (2019) suggested that higher WFPS (%) reduces N_2O emissions due to consumption and low
493 gas diffusivity. Similar results for simulating WFPS (%) by DNDC in multiple and
494 monoculture crops were reported in previous studies (e.g. Abdalla et al., 2014; Cui et al., 2014;
495 Li et al., 2017). The range of r^2 between simulated and observed values reported in these
496 previous studies was 0.1 to 0.6, compared to 0.4 to 0.5 found in this study. However, a previous
497 study found that the underestimation of water dynamics by the DNDC, in a similar studies in
498 North China plain, was due to the model uncertainty in estimating potential evapotranspiration
499 (Kröbel et al., 2010). To further improve the simulation of WFPS (%) for double cropping
500 system, the water module of DNDC needs to be further improved and any impact on the other
501 submodules of the model should be considered.

502 The DNDC underestimated the magnitude of daily soil N (exchangeable NH_4^+ and NO_3^-
503) concentrations. Similar findings were showed by Abdalla et al. (2014) for a reduced tillage-
504 cover crop experiment. The underestimation of WFPS (%) by DNDC, especially for the control

505 treatment, could be one of the reasons behind this underestimation of daily soil N. The presence
506 of two crops growing consecutively in the double cropping system increased the amount of C
507 and N turnover from crop residues and made it difficult for the model to correctly simulate
508 daily soil N. New features to quantify added C and N from crop residue are needed and the
509 algorithms for simulating these multiple cropping systems in the double cropping system need
510 to be improved.

511

512 **5 Conclusions**

513

514 In this study, the calibrated and evaluated DNDC model was able to effectively estimate
515 cumulative N₂O flux and grain yields from the summer maize-winter wheat double cropping
516 system. Conversely, the model generally underestimated daily soil N and WFPS (%) across all
517 the N management regimes. The high sensitivity of the DNDC model to rainfall, SOC and
518 temperature resulted in significant overestimation of N₂O peaks especially during the warm
519 wet season. The DNDC model is weak in simulating the control treatment. To further improve
520 the model's performance, further future studies are needed to identify and resolve the existing
521 problems especially with the control treatment.

522

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524

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529 **References**

530

531 Abdalla M, Wattenbach M, Smith P, Ambus P, Jones M, Williams M (2009a).

532 Application of the DNDC model to predict emissions of N₂O from Irish agriculture.

533 *Geoderma* 151: 327-337.

534 Abdalla M, Jones M, Smith P, Williams M (2009b). Nitrous oxide fluxes and

535 denitrification sensitivity to temperature in Irish pasture soils. *Soil Use Manag* 25:

536 376-388.

537 Abdalla M, Jones M, Ambus P, Williams M (2010). Emissions of nitrous oxide from

538 Irish arable soils: effects of tillage and reduced N input. *Nutr Cycl Agroecosyst* 86:
539 53-65.

540 Abdalla M, Rueangritsarakul K, Jones M, Osborne B, Helmy M, Roth B, Burke J,
541 Nolan P, Smith P, Williams M (2012). How effective is reduced tillage-cover crop
542 management in reducing N₂O fluxes from arable crop soils? *Water, Air Soil Poll* 223:
543 5155-5174.

544 Abdalla M, Hastings A, Helmy M, Prescher A, Osborne B, Lanigan G, Forristal D, Killi D,
545 Maratha P, Williams M, Rueangritsarakul K, Smith P, Nolan P, Jones NB
546 (2014). Assessing the combined use of reduced tillage and cover crops for mitigating
547 greenhouse gas emissions from arable ecosystem. *Geoderma* 223-225: 9-20.

548 Addiscott T M (1983). Kinetics and temperature relationships of mineralization and
549 nitrification in Rothamsted soils with differing histories. *Soil Sci* 34: 343-353.

550 Aulakh M, Rennie D, Paul E (1984). The influence of plant residues on denitrification
551 rates in conventional and zero-tilled soils. *Soil Sci Soc Am J* 48: 790-794.

552 Baggs EM, Blum H (2004). CH₄ oxidation and emissions of CH₄ and N₂O from *Lolium*
553 perenne swards under elevated atmospheric CO₂. *Soil Biol Biochem* 36: 713-723.

554 Bell MJ, Jones E, Smith J, Smith P, Yeluripati J, Augustin J, Juszczak R, Olejnik
555 J, Sommer M (2012). Simulation of soil nitrogen, nitrous oxide emissions and
556 mitigation scenarios at 3 European cropland sites using the ECOSSE model. *Nutr*
557 *Cycl Agroecosyst* 92:161-181.

558 Bowden WB, Bormann FH (1986). Transport and loss of nitrous oxide in soil water after
559 forest clear cutting. *Sci* 233: 867-869.

560 Cai Z, Swamoto T, Li C, Kang G, Boonjawat J, Mosier A, Wassmann R, Tsuruta
561 H (2003). Field validation of the DNDC-model for greenhouse gas emissions in East
562 Asian cropping systems. *Glob Biogeochem Cycl* 17 (4): 1107.

563 Chen X, Cui Z, Fan M, Vitousek P, Zhao M, Ma W., et al. (2014). Producing more
564 grain with lower environmental costs. *Nature* 514: 486-489.

565 Chu Z, Guo J, Zhao J (2017). Impacts of future climate change on agroclimatic resources
566 in Northeast China. *J Geograph Sci* 27(9): 1044-1058.

567 Collins S L, Knapp A K, Briggs J M, Steinauer E M (1998). Modulation of diversity
568 by grazing and mowing in native tallgrass prairie. *Sci* 280: 745-747.

569 Conrad R (1994). Compensation concentration as critical variable for regulating the flux of
570 trace gases between soil and atmosphere. *Biogeochem* 27: 155-170.

571 Conrad R (1996). Soil Microorganisms as Controllers of Atmospheric Trace Gases (H₂, CO,

572 CH₄, OCS, N₂O, and NO). *Microbiol Rev* 60: 609-640.

573 Cui F, Zheng X, Liu C, Wang K, Zhou Z, Deng J (2014). Assessing biogeochemical
574 effects and best management practice for a wheat-maize cropping system using the
575 DNDC model. *Biogeosci* 11: 91-107.

576 Cui Z, Yue S, Wang G, Zhang F, Chen X (2013). In-season root zone N management for
577 mitigating greenhouse gas emission and reactive N losses in intensive wheat
578 production. *Environ Sci Technol* 47 (11): 6015-6022.

579 Davidson EA, Matson PA, Vitousek PM, Riley R, Dunkin K, García-Méndez G, Maass JM
580 (1993). Processes regulating soil emissions of NO and N₂O in a seasonally dry
581 tropical forest. *Ecol* 74 (1):130-139

582 Deng J, Zhu B, Zhou Z, Zheng X, Li C, Wang T, Tan J (2011). Modeling nitrogen
583 loadings from agricultural soils in southwest China with modified DNDC. *J Geophys
584 Res* 116: G02020.

585 Ding W, Meng L, Yin Y, Cai Z, Zheng X (2007). CO₂ emission in an intensively
586 cultivated loam as affected by long-term application of organic manure and nitrogen
587 fertilizer. *Soil Biol Biochem* 39: 669-679.

588 Dobbie KE, Smith KA (2001). The effects of temperature, water filled pore space and land
589 use on N₂O emissions from an imperfectly drained gleysol. *Euro J Soil Sci* 52: 667-
590 673.

591 Ely A, Geall S, Song Y (2016). Sustainable maize production and consumption in China:
592 practices and politics in transition. *J Clean Prod* 134 (Part A): 259-268.

593 FAO (2017) FAOFAOSTAT Maize Production Data/China mainland (2017).
594 <http://faostat3.fao.org/>. (accessed on 24.10.19.).

595 Frohling SE, Mosier AR, Ojima DS, Li C, Parton WJ, Potter CS, Priesack E,
596 Stenger R, Haberbosch C, Dorsch P, Flessa H, Smith KA (1998). Comparisons of
597 N₂O emissions from soils at three temperate agricultural sites: simulations of year-
598 round measurements by four models. *Nutr Cycl Agroecosyst* 52: 77-105.

599 Galloway JN (1998). The global nitrogen cycle. Changes and consequences. *Environ
600 Poll* 102: 15-24.

601 Gilhespy S, Anthony S, Cardenas L, Chadwick D, del Prado A et al. (2014). First 20
602 years of DNDC (DeNitrification DeComposition): Model evolution. *Ecol Mod* 292:
603 51-62.

604 Giltrap DL, Li C, Saggart S (2010). DNDC: a process-based model of greenhouse gas
605 fluxes from agricultural soils. *Agric Ecosyst Environ* 136: 292-300.

606 Haas E, Klatt S, Fröhlich A, Kraft P, Werner C, Kiese R, Grote R, Breuer L,
607 Butterbach-Bahl K (2013). Landscape DNDC: a process model for simulation of
608 biosphere-atmosphere-hydrosphere exchange processes at site and regional scale.
609 *Landsc Ecol*, 28, 615-636.

610 Hastings AF, Wattenbach M, Eugster W, Li C, Buchmann N, Smith P (2010).
611 Uncertainty propagation in soil greenhouse gas emission models: An experiment
612 using the DNDC model and at the Oensingen cropland site. *Agric Ecosyst Environ*
613 136: 97-110.

614 Huang T, Gao B, Hu XK, Lu X, Well R, Christie P, Bakken LR, Ju XT (2014). Ammonia-
615 oxidation as an engine to generate nitrous oxide in an intensively managed calcareous
616 Fluvo aquic soil. *Sci Rep* 4:3950.

617 Hu L, Jian-jun Q, Li-gang W, Ming-yi X, Zhi-qiang L, Wei W (2012). Estimates of
618 N₂O emissions and mitigation potential from a spring maize field based on DNDC
619 model. *J Integ Agric* 11(12): 2067-2078.

620 Hu L, Ligang W, Jianzheng L, Maofang G, Jing Z, Jianfeng Z, Jianjun Q, Jia D, Li
621 C, Frolking S (2017). The development of China-DNDC and review of its
622 applications for sustaining Chinese agriculture. *Ecol Mod* 348: 1-13.

623 Jinguo Y, Zheng N, Chenli W (2006). Vegetation NPP distribution based on MODIS data
624 and CASA model-A case study of northern Hebei province. *Chin Geograph Sci* 16:
625 334-341.

626 Ju X T, Lu X, Gao ZL, Chen XP, Su F, Martin K, Volker R, Christie P, Zhang
627 FS (2011). Processes and factors controlling N₂O production in an intensively
628 managed low carbon calcareous soil under sub-humid monsoon conditions. *Environ*
629 *Poll* 159: 1007-1016.

630 Kesik M, Bruggemann N, Forkel R, Kiese R, Knoche R, Li C, Seufert G,
631 Simpson D, Butterbach-Bahl K (2006). Future scenarios of N₂O emissions from
632 European forest soils. *J Geophys Res* 111: G02018.

633 Khalil MAK, Rasmussen RA, Shearer MJ (2002). Atmospheric nitrous oxide:
634 patterns of global change during recent decades and centuries. *Chemosph* 47: 807-
635 821.

636 Khalil MI, Abdalla M, Lanigan G, Osborne B, Müller C (2016). Evaluation of
637 Parametric Limitations in Simulating Greenhouse Gas Fluxes from Irish Arable Soils
638 Using Three Process-Based Models. *Agric Sci* 7 (8): 503-520.

639 Kröbel R, Sun QP, Ingwersen J (2010). Modelling water dynamics with DNDC and

640 DAISY in a soil of north China plain: a comparative study. *Environ Mod Software*
641 25: 583-601.

642 Kuang W, Gao X, Gui D, Tenuta M, Flaten DN, Yin M, Zeng F (2018). Effects of fertilizer
643 and irrigation management on nitrous oxide emission from cotton fields in an
644 extremely arid region of north-western China. *Field Crops Res* 229: 17-26.

645 Kuang W, Gao X, Tenuta M, Gui D, Zeng F (2019). Relationship between soil profile
646 accumulation and surface emission of N₂O: effects of soil moisture and fertilizer
647 nitrogen. *Nutr Fert soils* 55 (2):97-107.

648 Laik R, Sharma ., Idris M, Singh AK, Singh SS, Bhatt BP, Saharawat Y,
649 Humphreys E, Ladha JK (2014). Integration of conservation agriculture with best
650 management practices for improving system performance of the rice-wheat rotation
651 in the Eastern Indo-Gangetic Plains of India. *Agric Ecosyst Environ* 195: 68-82.

652 Lamers M, Ingwersen J, Streck T (2007). Modelling N₂O emission from a forest upland
653 soil: a procedure for an automatic calibration of the biogeochemical model Forest-
654 DNDC. *Ecol Mod* 205: 52-58.

655 Li C, Frolking S, Frolking TA (1992). A model of nitrous oxide evolution from soil
656 driven by rainfall events.1. Model structure and sensitivity. *J Geophys Res Atmos* 97:
657 9759-9776.

658 Li C, Frolking S, Harris, R (1994). Modelling carbon biogeochemistry in agricultural
659 soils. *Glob Biogeochem Cycl* 8: 237-254.

660 Li C, Aber J, Stange F, Butterbach-Bahal K, Papen H (2000). A process-oriented model
661 of N₂O and NO emissions from forest soils: 1. Model development. *J Geophys Res*
662 105: 4369-4384.

663 Li C, Zhuang Y, Cao M, Crill P, Dai Z, Frolking S, Moore B, Salas W, Song W,
664 Wang X (2001). Comparing a process-based agro ecosystem model to the IPCC
665 methodology for developing a national inventory of N₂O emissions from arable lands
666 in China. *Nutr Cycl Agroecosyst* 60: 1-3.

667 Li H, Qiu J, Wang L, Xu M, Liu Z, Wang W (2012). Estimates of N₂O Emissions and
668 Mitigation Potential from a Spring Maize Field Based on DNDC Model. *J Integ Agric*
669 11(12): 2067-2078.

670 Li J, Van Bueren EL, Jiggins J, Leeuwis C (2012). ‘Farmers’ adoption of maize (*Zea*
671 *mays* L.) hybrids and the persistence of landraces in Southwest China: implications
672 for policy and breeding. *Gen Res Crop Evol* 59: 1147-1160.

673 Li H, Ligang W, Jianzheng L, Maofang G, Jing Z, Jianfeng Z, Jianjun Q, Jia D,

674 Li C, Frolking S (2017). The development of China-DNDC and review of its
675 applications for sustaining Chinese agriculture. *Ecol Mod* 348: 1-13.

676 Li H, Wang L, Qiu J, Li C, Gao M, Gao C (2014). Calibration of DNDC model for
677 nitrate leaching from an intensively cultivated region of Northern China. *Geoderma*
678 223-225: 108-118.

679 Li KY, Zhao YY, Yuan XL, Zhao HB, Wang ZH, Li HX, Malhi SS (2012). Comparison of
680 Factors Affecting Soil Nitrate Nitrogen and Ammonium Nitrogen Extraction. *Comm*
681 *Soil Sci Plant Anal* 43:3, 571-588.

682 Ludwig B, Jäger N, Priesack E, Flessa H (2011). Application of the DNDC model to predict
683 N₂O emissions from sandy arable soils with differing fertilization in a long - term
684 experiment. *J Plant Nutr Soil Sci* 174(3): 350-358.

685 Mathieu O, Henault C, Leveque J, Baujard E, Milloux MJ, Andreux F (2006). Quantifying
686 the contribution of nitrification and denitrification to the nitrous oxide flux using 15N
687 tracers. *Environ Poll* 144: 933-940.

688 Nash JE, Sutcliffe JV (1970). River flow forecasting through conceptual models - part 1: a
689 discussion of principles. *J Hydrol* 10: 282-290.

690 Norse D (2011). SAIN. Improved Nutrient Management in Agriculture - a Neglected
691 Opportunity for China's Low Carbon Growth Path, Policy Brief No. 1. Sustainable
692 Agricultural Innovation Network (2011). [http://www.eu-china.net/upload/
693 pdf/materialien/11-02-11_PolicyBriefNo1updatedfinal.pdf](http://www.eu-china.net/upload/pdf/materialien/11-02-11_PolicyBriefNo1updatedfinal.pdf). (Accessed on 2.09.18).

694 Parkin TB, Kaspar TC (2006). Nitrous oxide emissions from corn-soybean systems in the
695 Midwest. *J Environ Qual* 35: 1496-1506

696 Parkin TB (2008). effect of sampling frequency on estimates of cumulative nitrous oxide
697 emissions. *J Environ Qual* 37: 1390-1395.

698 Patra AK, Abbadie L, Clays A, Degrange V, Grayston S, Guillaumaud N, Loiseau
699 P, Louault F, Mahmood S, Nazaret S, Philippot L, Poly F, Prosser JI, Le Roux X
700 (2006). Effects of management regime and plant species on the enzyme activity and
701 genetic structure of N-fixing, denitrifying and nitrifying bacterial communities in
702 grassland soils. *Environ Microbiol* 8: 1005-1016.

703 Powlson D, Norse D, Lu Y (2018). SAIN. Agricultural development in China -
704 environmental impacts, sustainability issues and policy implications assessed through
705 China-UK projects under SAIN (UK-China Sustainable Agriculture Innovation
706 Network), 2008 – 2017.

707 file:///G:/DNDC%20trail/Corrections%20from%20the%20Chinese/SAIN,
708 %202018.pdf. (Accessedd on 24.10.18).

709 Roelandt C, Van Wesemael B, Rounsevell M (2005). Estimating annual N₂O emissions from
710 agricultural soils in temperate climate. *Glob Chan Biol* 11: 1701-1711.

711 Sarkodie-Addo, J., Lee, H. C., Baggs, E. M. (2003). Nitrous oxide emissions after application
712 of inorganic fertilizer and incorporation of green manure residues. *Soil Use Manag*
713 19: 33-339.

714 Smith J, Smith P (2007). *Environmental modelling: an introduction* (pp. 1-178).
715 Oxford University Press, Oxford, UK.

716 Smith WN, Grant BB, Desjardins RL, Worth D, Li C, Boles SH, Huffman EC
717 (2010). A tool to link agricultural activity data with the DNDC model to estimate
718 GHG emissions factor in Canada. *Agric Ecosyst Environ* 136: 301-309.

719 Song X, Liu M, Ju X, Gao B, Su F, Chen X, Rees RM (2018). Nitrous oxide emissions
720 increase exponentially when optimum nitrogen fertilizer rates are exceeded in the North
721 China plain. *Environ Sci Tech* 52 (21): 12504-12513.

722 Song X, Ju X, Topp CFE, Rees RM (2019). Oxygen regulates nitrous oxide production
723 directly in agricultural soils. *Environ Sci Tech* 53 (21): 12539-12547.

724 Tonitto C, David MB, Drinkwater LE, Li C (2007). Application of the DNDC model to
725 tile-drained illinois agroecosystems: model calibration, validation, and sensitivity
726 analysis. *Nutr Cycl Agroecosyst* 78 (1): 51-63.

727 Ussiri D, Lal R (2012). *Soil Emission of N₂O and its Mitigation*. Dordrecht: Springer.

728 Uzoma KC, Smith W, Grant B, Desjardins RL, Gao X, Hanis K, Tenuta M, Goglio
729 P, Li C (2015) Assessing the effects of agricultural management on nitrous oxide
730 emissions using flux measurements and the DNDC model. *Agric Ecosyst Environ*
731 206: 71-83

732 Van Oijen M, Cameron DR, Butterbach-Bahl K, Farahbakhshazad N, Jansson PE,
733 Kiese R, Rahn KH, Werner C, Yeluripati JB (2011). A Bayesian framework for
734 model calibration, comparison and analysis: Application to four models for the
735 biogeochemistry of a Norway spruce forest. *Agric ForMeteo* 151: 1609-1621.

736 Wagner-Riddle C, Furon NL, Mclaughlin IL, Barbeau J, Jayasundara S, Parkin G,
737 von Bertoldi P, Warland J (2007). Intensive measurement of nitrous oxide emissions
738 from a corn- soybean-wheat rotation under two contrasting management systems over
739 5 years. *Glob Chan Biol* 13: 1722-1736.

740 Wattenbach M, et al (2010). The carbon balance of European croplands: A cross-site

741 comparison of simulation models. *Agric Ecosyst Environ* 139: 419-453.

742 Wennman P, Katterer T (2006). Effects of moisture and temperature on carbon and nitrogen
743 mineralisation in mine tailing mixed with sewage sludge. *J Environ Qual* 35: 1135-
744 1141.

745 Willams EJ, Hutchinson GL, Feshsenfeld FC (1992). NO_x and N₂O emissions from soil.
746 *Glob Biogeochem Cycl* 6: 351-388.

747 Wolf B, Zheng XH, Brueggemann N, et al (2010). Grazing-induced reduction of natural
748 nitrous oxide release from continental steppe. *Nature* 464: 881-884.

749 Xiong Z Q, Xing GX, Tsuruta H, Shen GY, Shi SL, Du LJ (2002). Measurement
750 of nitrous oxide emissions from two rice-based cropping systems in China. *Nutr Cycl*
751 *Agroecosyst* 64: 125-133.

752 Yang JM, Yang JY, Liu S, Hoogenboom G (2014). An evaluation of the statistical
753 methods for testing the performance of crop models with observed data. *Agric Syst*
754 127: 81-89.

755 Yan G, Yao Z, Zheng X, Liu C (2015). Characteristics of Annual Nitrous and Nitric Oxide
756 Emissions from Major Cereal Crops in the North China Plain under Alternative
757 Fertilizer Management. *Agric Ecosyst Environ* 207: 67-78.

758 Yue Q, Cheng K, Ogle S, Hillier J, Smith P, Abdalla M, Ledo A, Sun J, Pan G
759 (2018). Evaluation of four modelling approaches to estimate nitrous oxide emissions
760 in China's cropland. *Sci Tot Environ* 652: 1279-1289.

761 Zhang Y, Li C, Zhou X, Moore III B (2002). A simulation model linking crop growth and
762 soil biogeochemistry for sustainable agriculture. *Ecol Mod* 151: 75-108.

763 Zhang Y, Niu H (2016). The development of the DNDC plant growth sub-model and the
764 application of DNDC in agriculture: A review. *Agric Ecosyst Environ* 230: 271-282.

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775 **Tables**

776

777 **Table 1** Crop parameters used to calibrate the DNDC model for grain yield in each cropping season and simulated and observed grain yields.

778

Cropping season/ parameter	Grain	Leaf	Stem	Root	Simulated yield (t ha ⁻¹)	Observed yield (t ha ⁻¹)
<i>Summer maize 2012</i>						
Maximum biomass production (kg C ha ⁻¹ y ⁻¹)	3850	1694	1694	462	3.9	3.9
Biomass fraction	0.5	0.22	0.22	0.06		
Biomass C/N ratio	50	80	80	80		
Thermal degree days	2550					
Water demand (g water/g DM)	150					
Optimum temperature (°C)	30					
<i>Winter wheat 2012-2013</i>						
Maximum biomass production (kg C ha ⁻¹ y ⁻¹)	3300	1732	1732	1485	3.0	3.0
Biomass fraction	0.4	0.21	0.21	0.18		
Biomass C/N ratio	40	95	95	95		
Thermal degree days	1300					
Water demand (g water/g DM)	200					
Optimum temperature (°C)	22					
<i>Summer maize 2013</i>						
Maximum biomass production (kg C ha ⁻¹ y ⁻¹)	3550	1562	1562	462	3.5	3.5
Biomass fraction	0.5	0.22	0.22	0.06		
Biomass C/N ratio	50	80	80	80		
Thermal degree days	2550					
Water demand (g water/g DM)	150					
Optimum temperature (°C)	30					
<i>Winter wheat 2013-2014</i>						
Maximum biomass production (kg C ha ⁻¹ y ⁻¹)	3300	1540	1540	953	2.8	2.8
Biomass fraction	0.45	0.21	0.21	0.13		
Biomass C/N ratio	40	95	95	95		
Thermal degree days	1300					
Water demand (g water/g DM)	200					
Optimum temperature (°C)	22					

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Table 2 Statistical evaluations of simulated daily soil temperature, WFPS, nitrate and cumulative N₂O fluxes compared with the observed values under different N management of summer maize -winter wheat double cropping system from 2012 to 2014.

Treatment/parameter	Observed	Simulated	RD (%)	RMSE	nRMSE (%)	EF	d
Control							
<i>Average daily soil temperature (°C)</i>	16.3	20.0	23	4.1	25	0.89	0.89
<i>Average daily WFPS (%)</i>	57.0	13.6	-76	42	74	<0	0.40
<i>Average daily soil N (kg N ha⁻¹)</i>	1.1	0.1	-91	0.54	4	0.58	0.58
<i>N₂O emissions</i>	1.1	1.5 (1.8)*	42	0.55	4	<0	0.10
Conventional N							
<i>Average daily soil temperature (°C)</i>	16.3	20.1	23	4.2	26	0.79	0.89
<i>Average daily WFPS (%)</i>	54.7	20.7	-62	12.9	24	<0	0.43
<i>Average daily soil N (kg N ha⁻¹)</i>	87.7	69.5	-21	2.39	3	0.11	0.75
<i>N₂O emissions</i>	12.0	10.4 (5.5)	-13	2.59	16	<0	0.50
Optimal N							
<i>Average daily soil temperature (°C)</i>	16.3	20.0	23	4.1	25	0.96	0.97
<i>Average daily WFPS (%)</i>	55.0	20.2	-63	37.4	67	0.10	0.51
<i>Average daily soil N (kg N ha⁻¹)</i>	49.7	28.6	-42	1.32	2	<0	0.57
<i>N₂O emissions</i>	6.9	7.9 (3.5)	16	1.9	20	<0	0.29
1.3*Optimal N							
<i>Average daily soil temperature (°C)</i>	16.3	20.0	23	4.1	25	0.96	0.97
<i>Average daily WFPS (%)</i>	55.0	20.1	-63	37.0	67	0.10	0.75
<i>Average daily soil N (kg N ha⁻¹)</i>	6.3	5.1	-19	0.27	4	0.02	0.74
<i>N₂O emissions</i>	8.6	9.5 (4.6)	10	2.18	20	<0	0.29

* The values between brackets represent the model results before calibration.

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802 **Table 3** Comparisons between the DNDC- simulated and observed annual grain yields (t ha⁻¹) (2012-2014) of the summer maize - winter wheat double cropping system
 803 before and after the DNDC model calibration.

Treatment	Grown seasonal crop	Season/ Year	Observed yield	Simulated yield (before)	Simulated yield (after)	RD (%; before)	RD (%; after)
Control	Summer maize	2012	6.7	0.2	4.8	-97	-30
	Summer maize	2013	5.2	0.4	3.0	-92	-40
Conventional N	Summer maize	2012	10.2	12.0	9.8	18	-5
	Summer maize	2013	9.5	11.4	9.0	20	-5
Optimal N	Summer maize	2012	9.5	10.5	9.8	11	7
	Summer maize	2013	9.7	10.0	9.0	03	-7
1.3* Optimal N	Summer maize	2012	10.4	11.1	9.7	07	-7
	Summer maize	2013	9.5	10.0	8.9	05	-6
Control	Winter wheat	2013	2.3	0.3	1.1	-87	-50
	Winter wheat	2014	2.3	0.4	0.9	-83	-60
Conventional N	Winter wheat	2013	8.2	13.0	8.0	59	-2
	Winter wheat	2014	7.9	5.8	6.3	-27	-21
Optimal N	Winter wheat	2013	8.0	8.8	8.0	11	0
	Winter wheat	2014	7.8	4.5	8.3	-42	6
1.3* Optimal N	Winter wheat	2013	8.0	11.3	8.0	41	0
	Winter wheat	2014	8.1	5.1	8.2	-37	2

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818 **Figure captions**

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820 **Fig. 1** Sensitivity analysis of the DNDC model to changes in the input parameters (i.e. daily precipitation, daily air temperature, soil organic C
821 (SOC), applied N fertilizer and water irrigation).

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823 **Fig. 2** Comparisons between DNDC- model-simulated (red lines) and field observed (●) daily N₂O fluxes from the control (a), conventional N
824 (b), optimal N (c), and 1.3*optimal N (d) fertilizer application rate over the experiment period of the maize-wheat double cropping system
825 (2012-2014). Black arrows show the date of N fertilizer application and blue arrows show the date of water irrigation. (Error bars for observed
826 values are ± standard error).

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828 **Fig. 3** Comparisons between the DNDC-model- simulated (line) and field observed (●) soil nitrate plus ammonium (kg N ha⁻¹) at 0-20cm depth
829 from the control (a; $r^2 = 0.15$), conventional (b; $r^2 = 0.17$), optimal N (c; $r^2 = 0.15$) and 1.3*optimal N (d; $r^2 = 0.11$). Arrows show times of
830 fertilizer application. (Error bars for observed values are ± standard error).

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832 **Fig. 4** Comparisons between the DNDC- model- simulated and field observed daily soil temperature (°C) at 0-5cm depth; for control (a),
833 conventional N (b), optimal N (c) and 1.3* optimal N (d).

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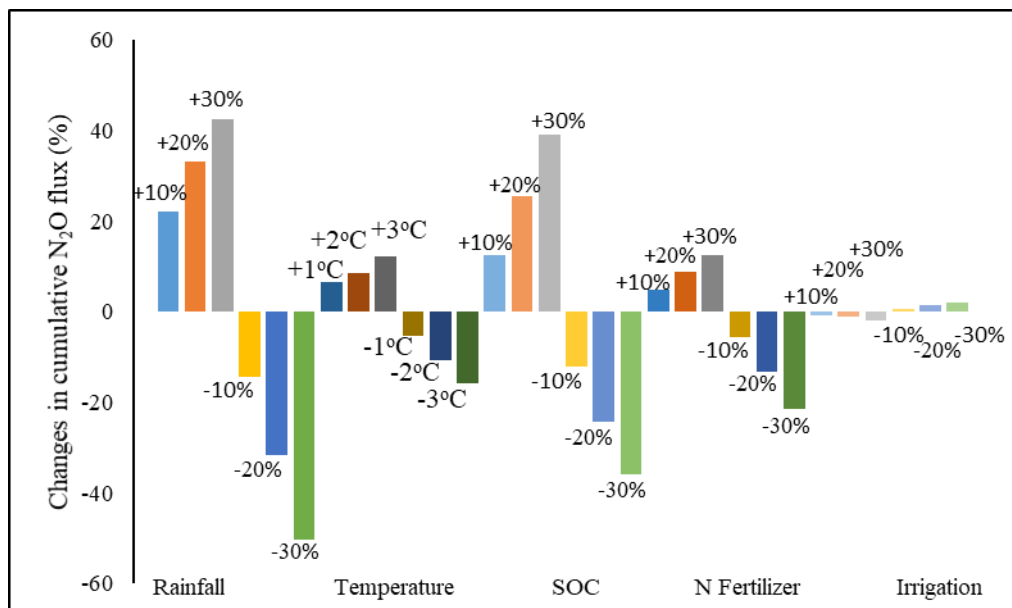
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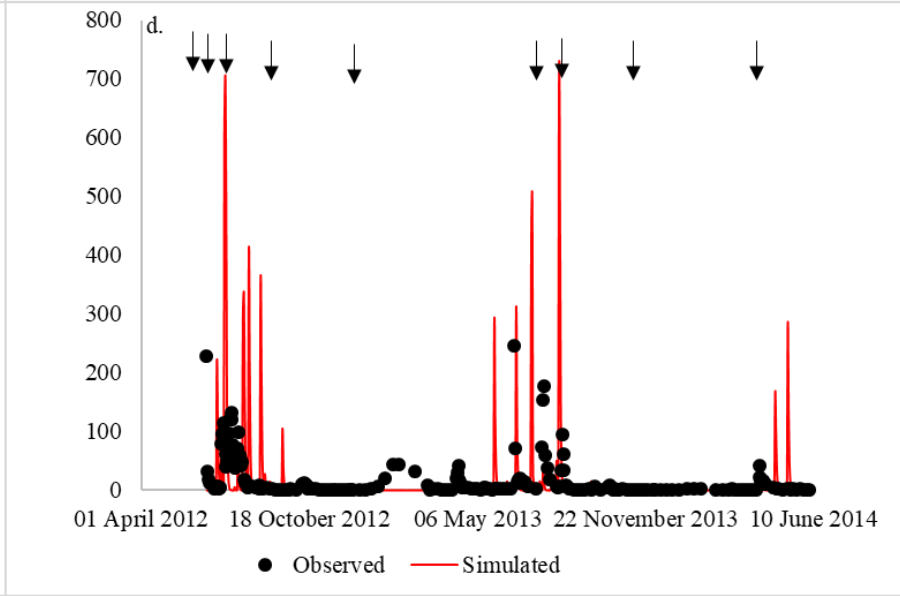
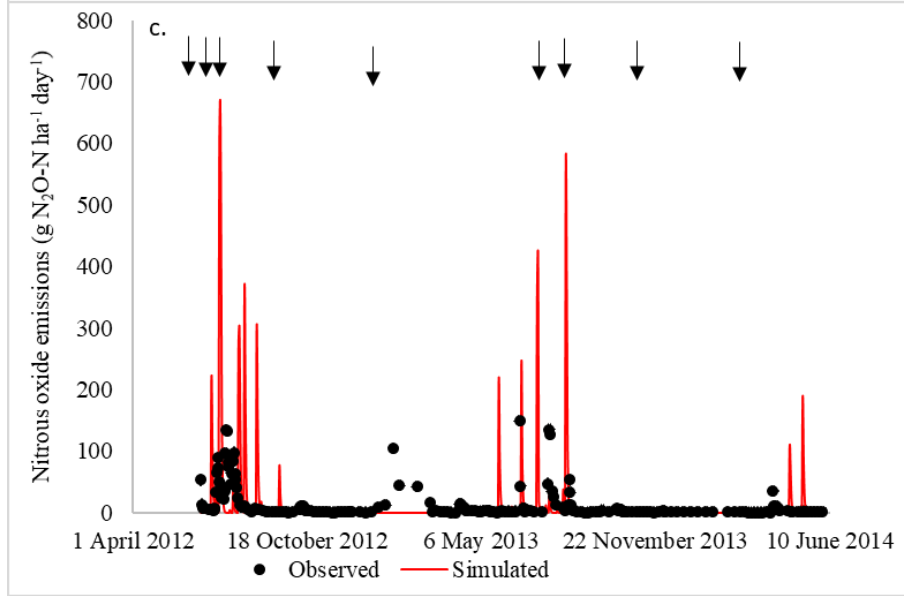
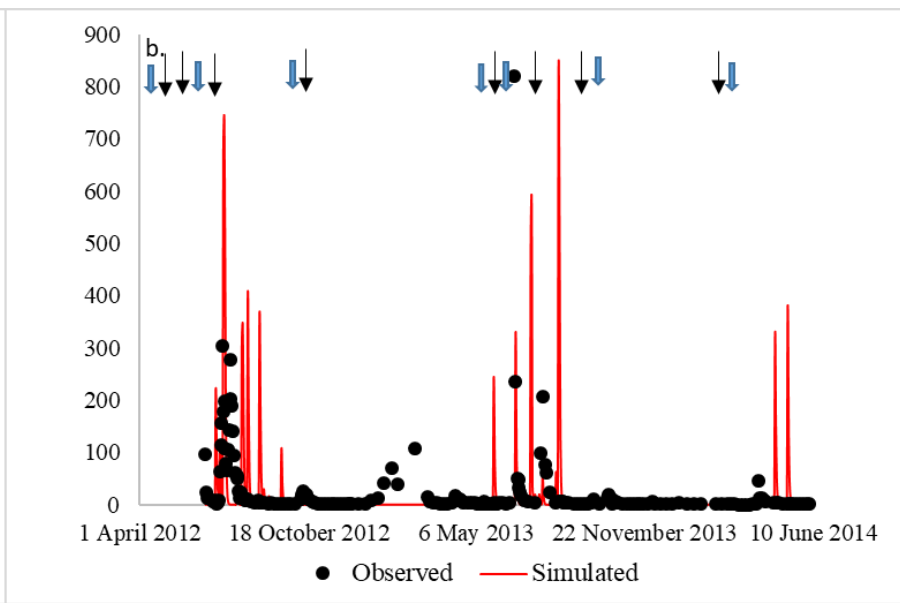
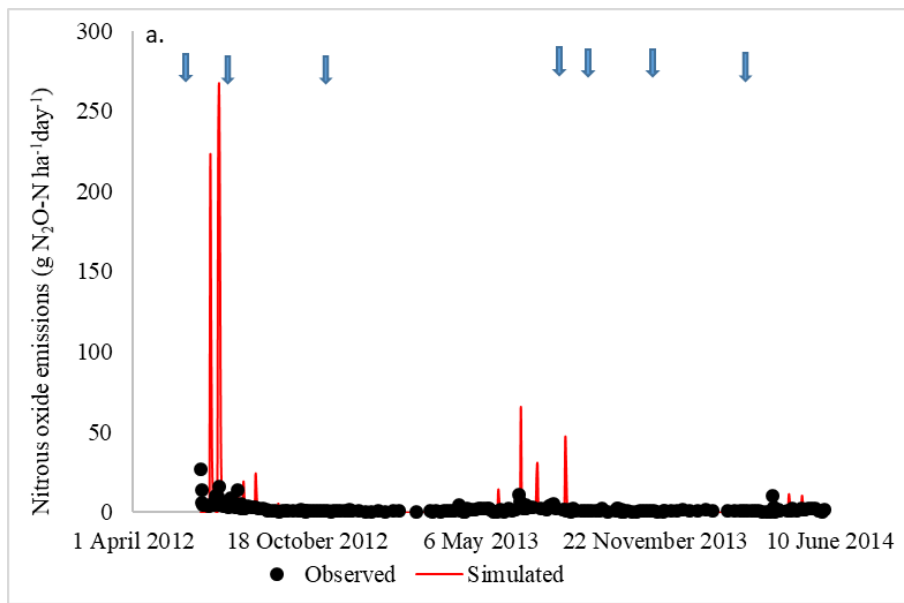
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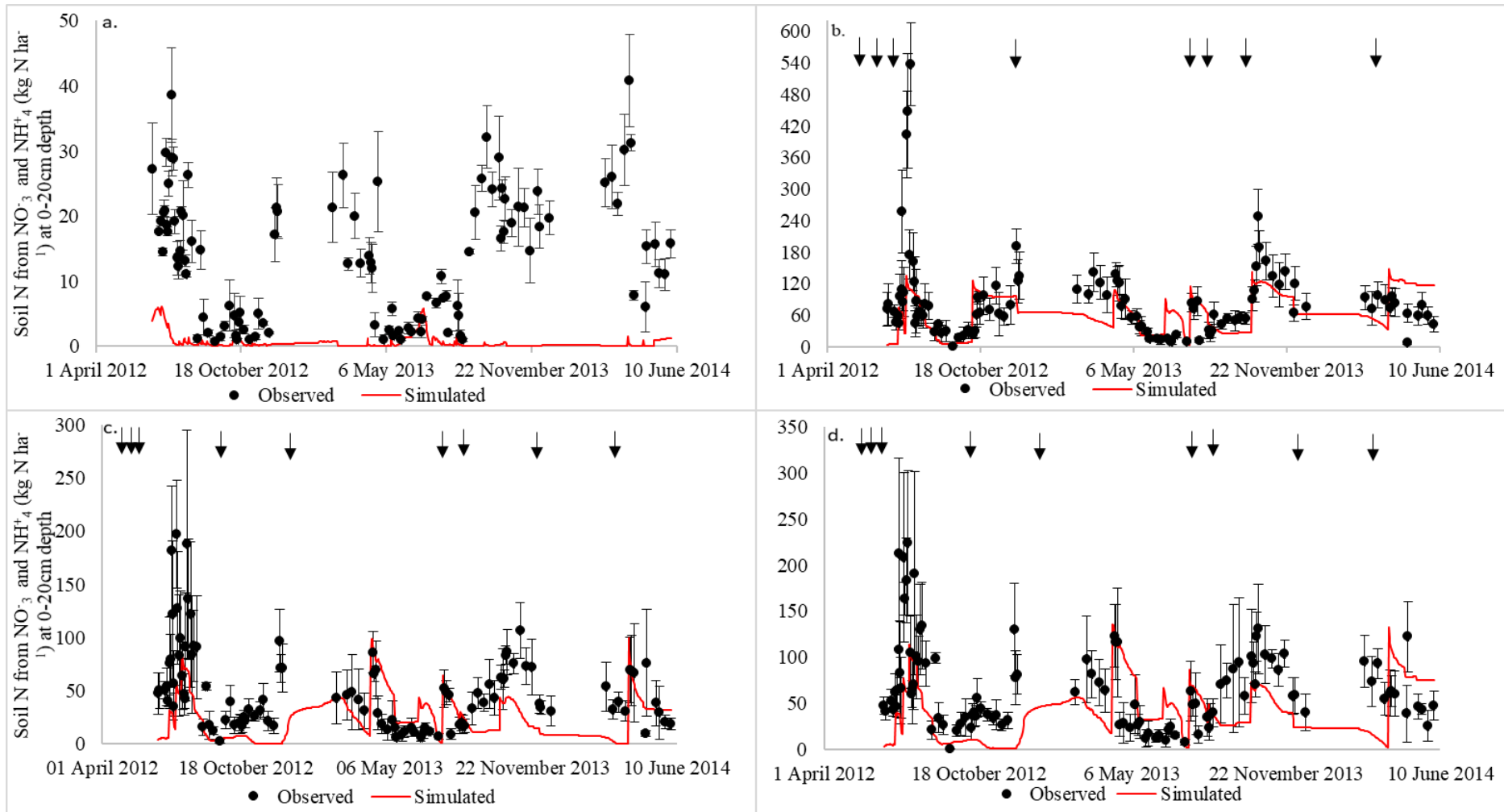
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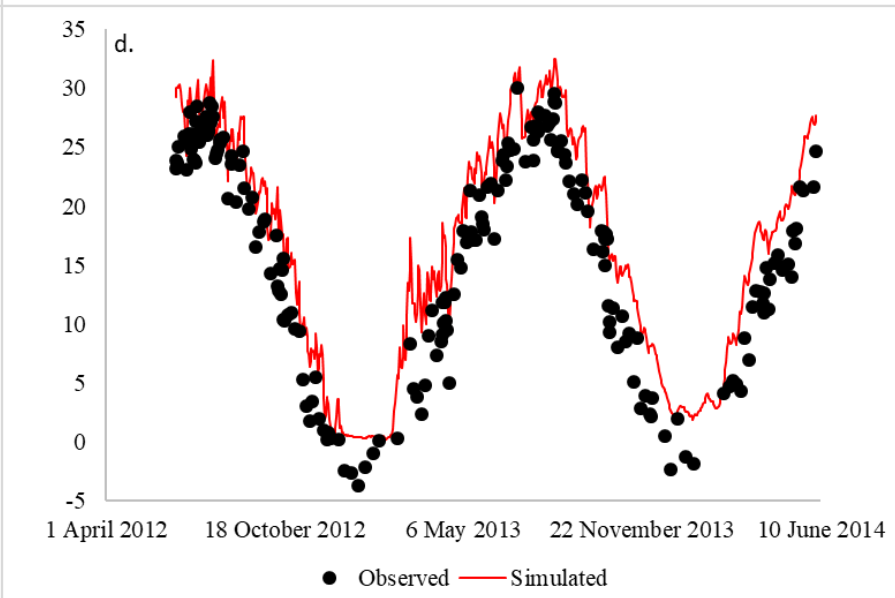
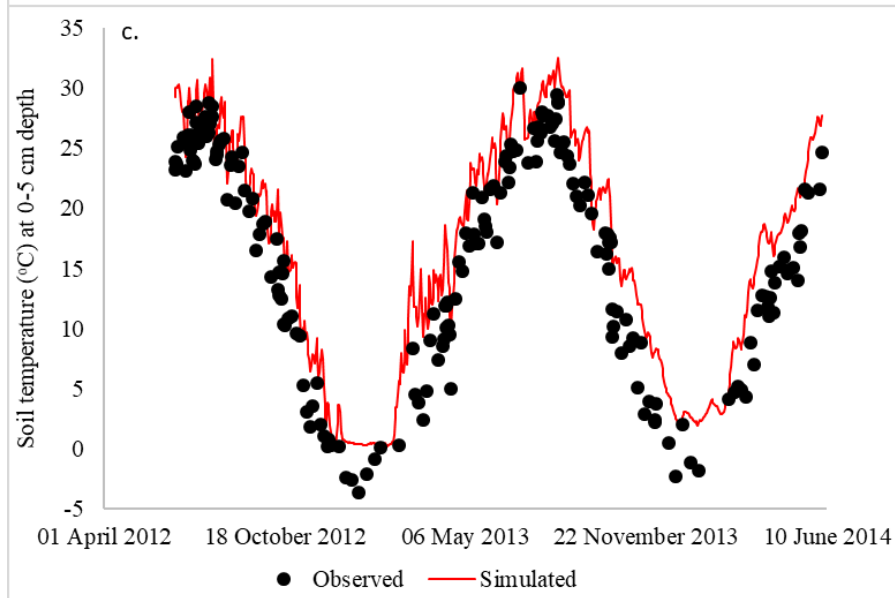
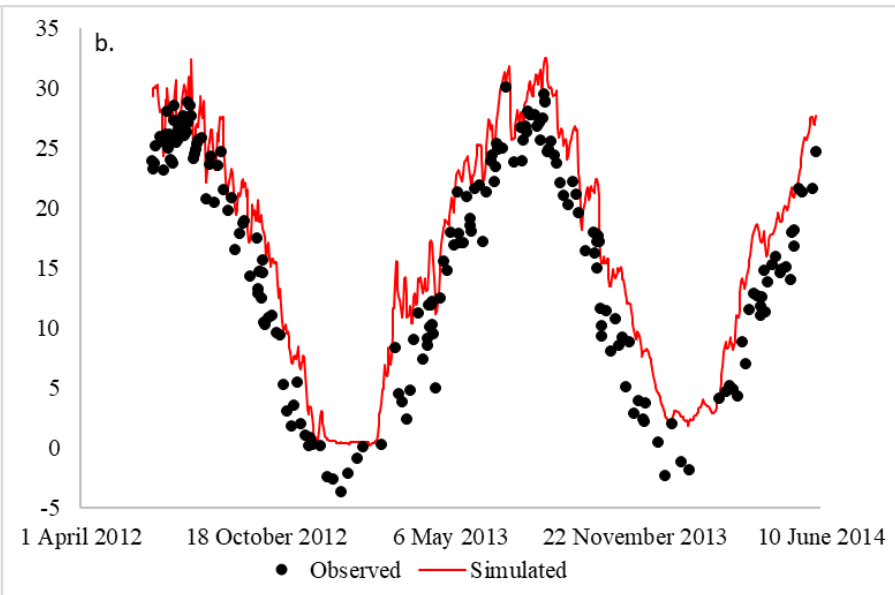
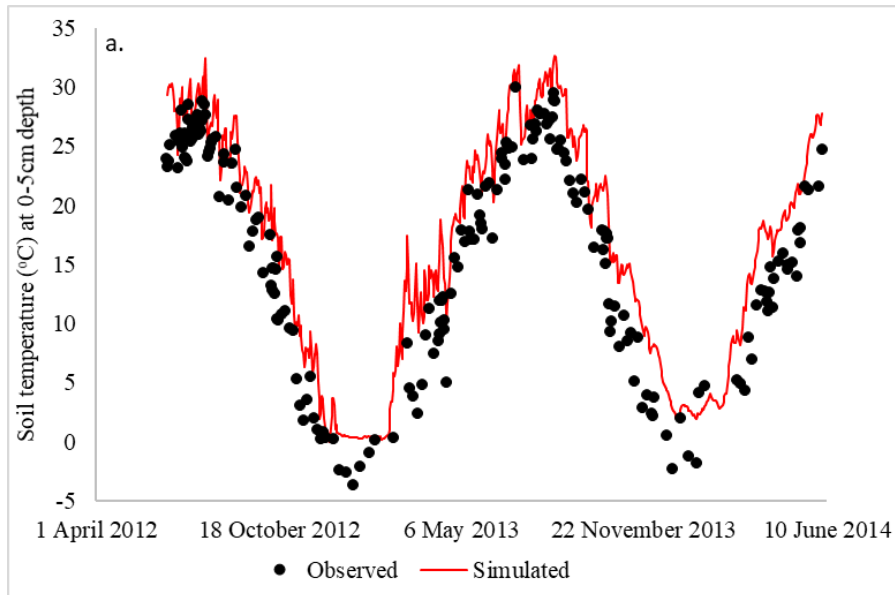


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858 **Supplementary Materials**

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860 **Table S1** Nitrogen fertilizer application rates (kg N ha⁻¹) and irrigation (mm) at the different N fertilizer management during the experimental period 2012-2014

Growing season	Date	Control	Conventional N	Optimal N	1.3*optimal N	0.7*optimal N	Irrigation rate	
2012 maize	17 June	0	-	-	-	-	90	861
	3 July	0	100 ^a	45 ^a	59 ^a	32 ^a	-	862
	13 July	0	150 ^b	69 ^b	89 ^b	48 ^b	-	863
	21 July	0	0	58 ^a	75 ^a	40 ^a	-	864
	Total	0	250	172	223	120	90	865
2012-2013 wheat	8 Oct. 2012	0	150 ^c	50 ^c	65 ^c	35 ^c	-	866
	5 Dec. 2012	0	0	0	0	0	75	867
	10 Apr. 2013	0	150 ^b	139 ^b	181 ^b	97 ^b	70	868
	13 May 2013	0	0	0	0	0	90	869
	Total	0	300	189	246	132	235	870
2013 maize	16 June	0	100 ^c	45 ^c	59 ^c	32 ^c	-	871
	18 June	0	-	-	-	-	75	872
	19 July	0	150 ^b	90 ^b	117 ^b	63 ^b	-	873
	13 August	0	0	30 ^b	39 ^b	21 ^b	-	874
	Total	0	250	165	215	116	75	875
2013-2013 wheat	7 Oct. 2013	0	150 ^c	50 ^c	65 ^c	35 ^c	-	876
	1 Dec. 2013	0	0	0	0	0	75	877
	4 Apr. 2014	0	150 ^b	127	165 ^b	89 ^b	90	878
	Total	0	300	177	230	124	165	879

885 Letters a-c represent the N application method: ^a= Band application followed by soil covering; ^b= Surface broadcast; ^c= incorporating surface applied N into soil.

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893 **Supplementary Figures**

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895 **Figure captions**

896 **Fig. S1** Average air temperature (°C) and daily precipitation (mm) at the experimental site during the study period of 2012-2014.

897

898 **Fig. S2:** A 1:1 relationship between the DNDC simulated and field observed cumulative N₂O emissions from the maize-wheat double cropping
899 system ($y = 0.99x$ and $r^2 = 0.91$).

900

901 **Fig. S3:** 1:1 relationships between DNDC-simulated and field observed grain yields; for maize/wheat combination (a; $r^2 = 0.91$), maize (b; $r^2 =$
902 0.89) and wheat (c; $r^2 = 0.92$).

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904 **Fig. S4:** 1:1 relationships between daily DNDC-simulated and field observed water filled pore space (WFPS; %) at 0-20 cm depth; for control (a;
905 $r^2 = 0.30$), conventional N (b; $r^2 = 0.37$), optimal N (c; $r^2 = 0.31$) and 1.3* optimal N (d; $r^2 = 0.37$). (Error bars for observed values are \pm standard
906 error).

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908 **Fig. S5:** A 1:1 relationship between the DNDC simulated and field observed cumulative soil N for the maize-wheat double cropping system ($y =$
909 $0.74x$; $r^2 = 0.97$).

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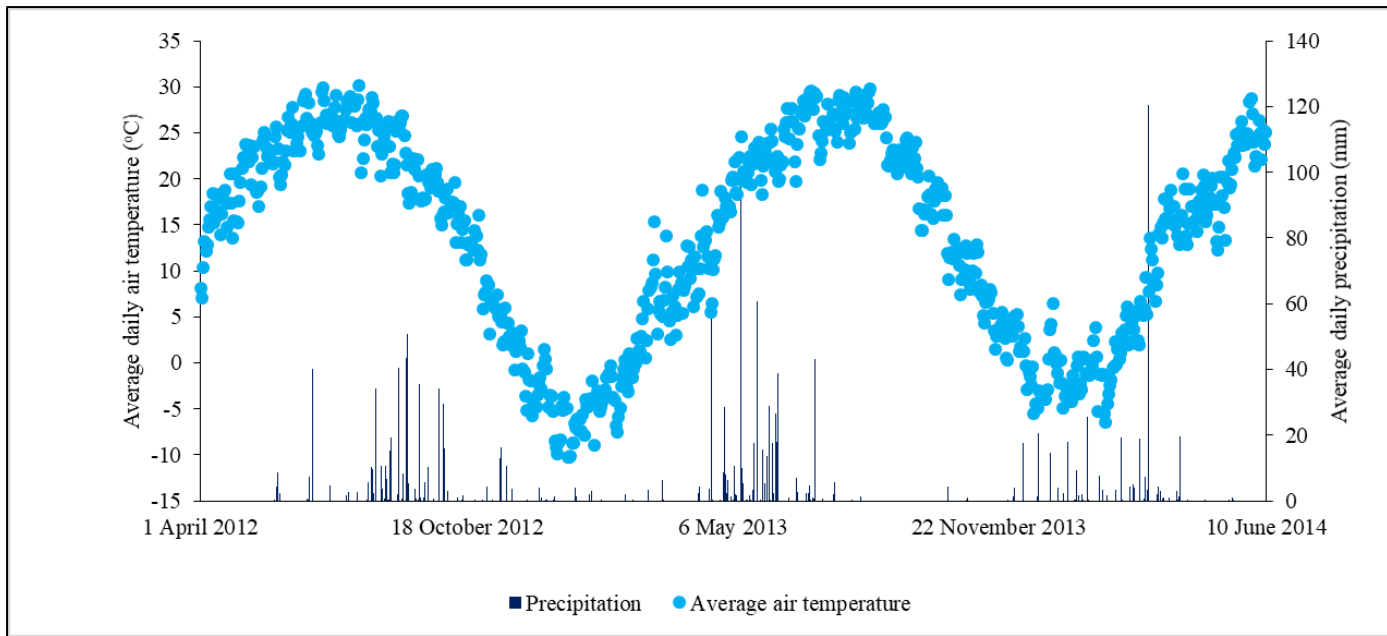
911 **Fig. S6:** 1:1 relationships between daily DNDC-simulated and field observed soil temperature (°C) at 0-5 cm depth; for control (a; $r^2 = 0.89$),
912 conventional N (b; $r^2 = 0.88$), optimal N (c; $r^2 = 0.88$) and 1.3* optimal N (d; $r^2 = 0.88$).

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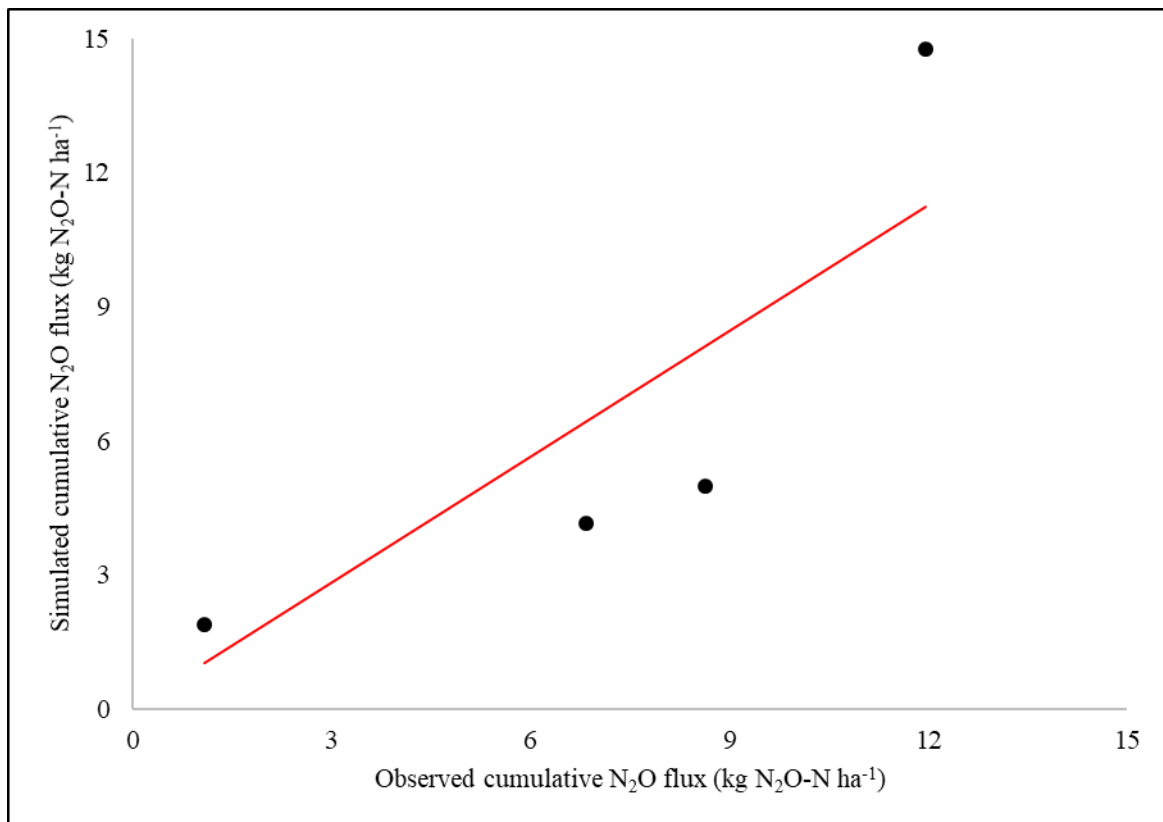
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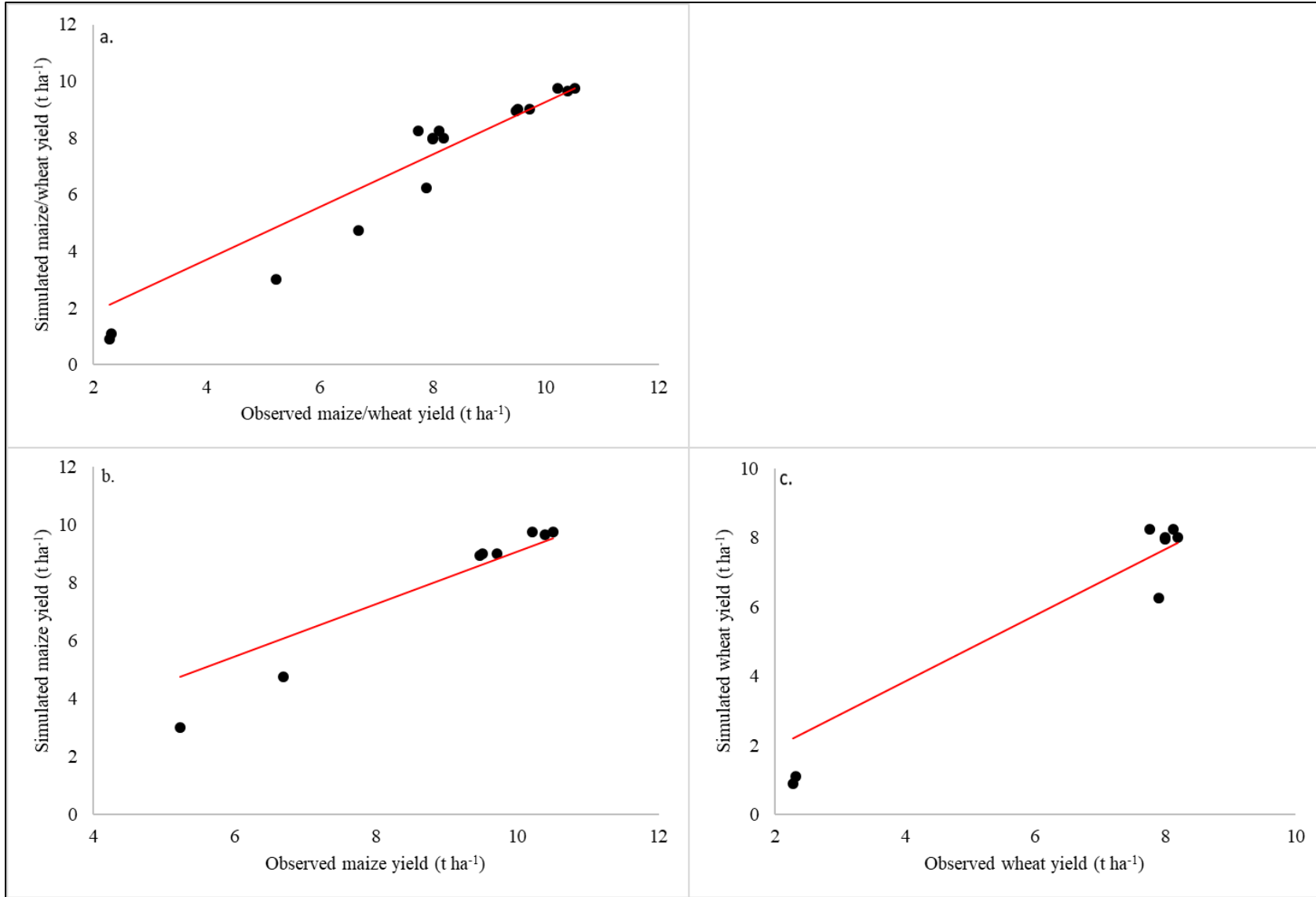
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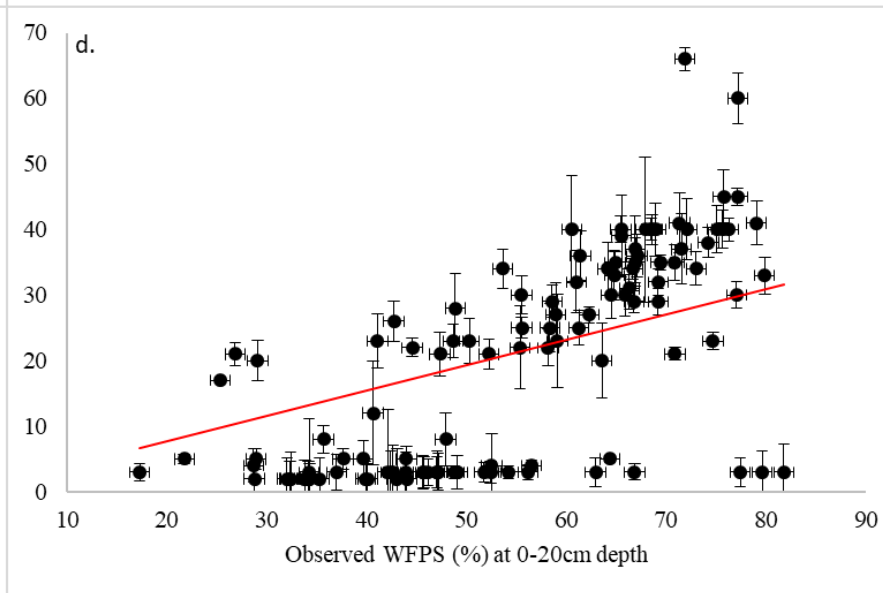
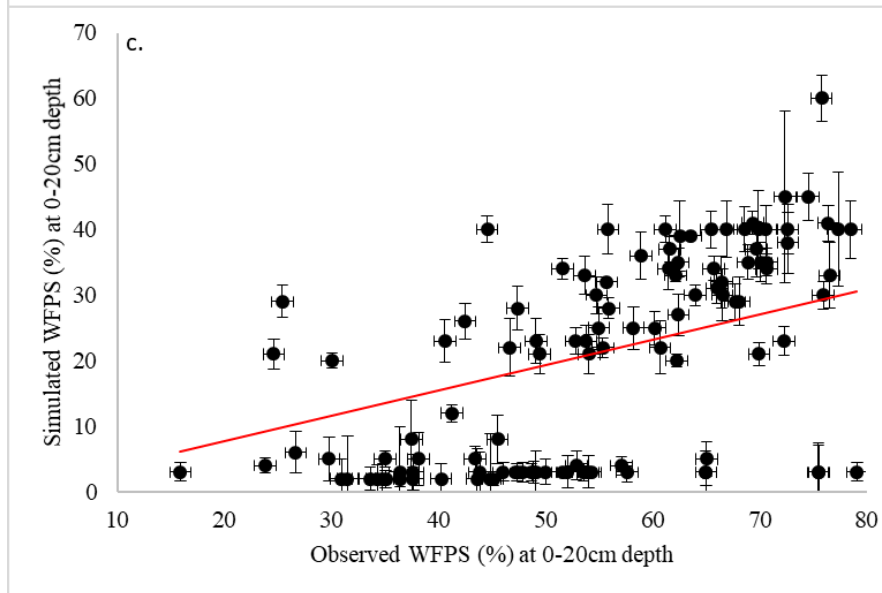
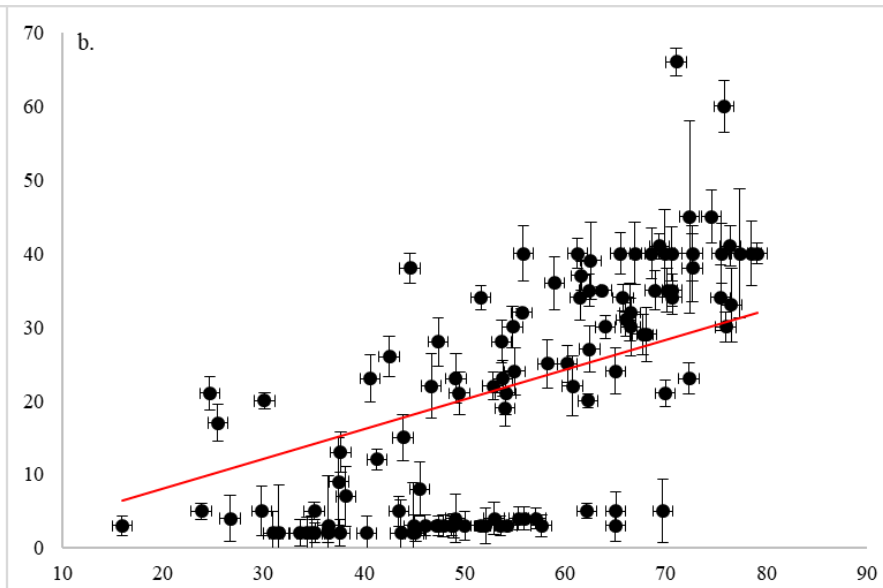
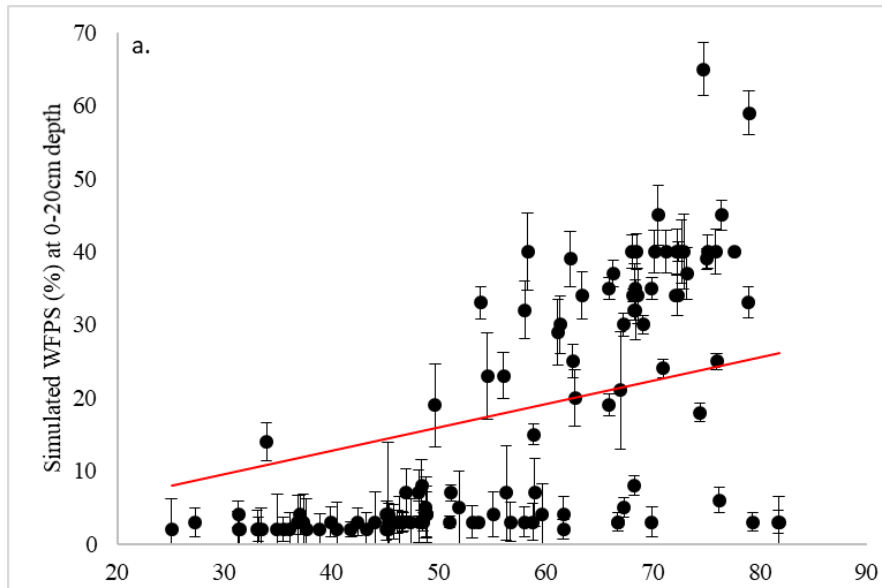
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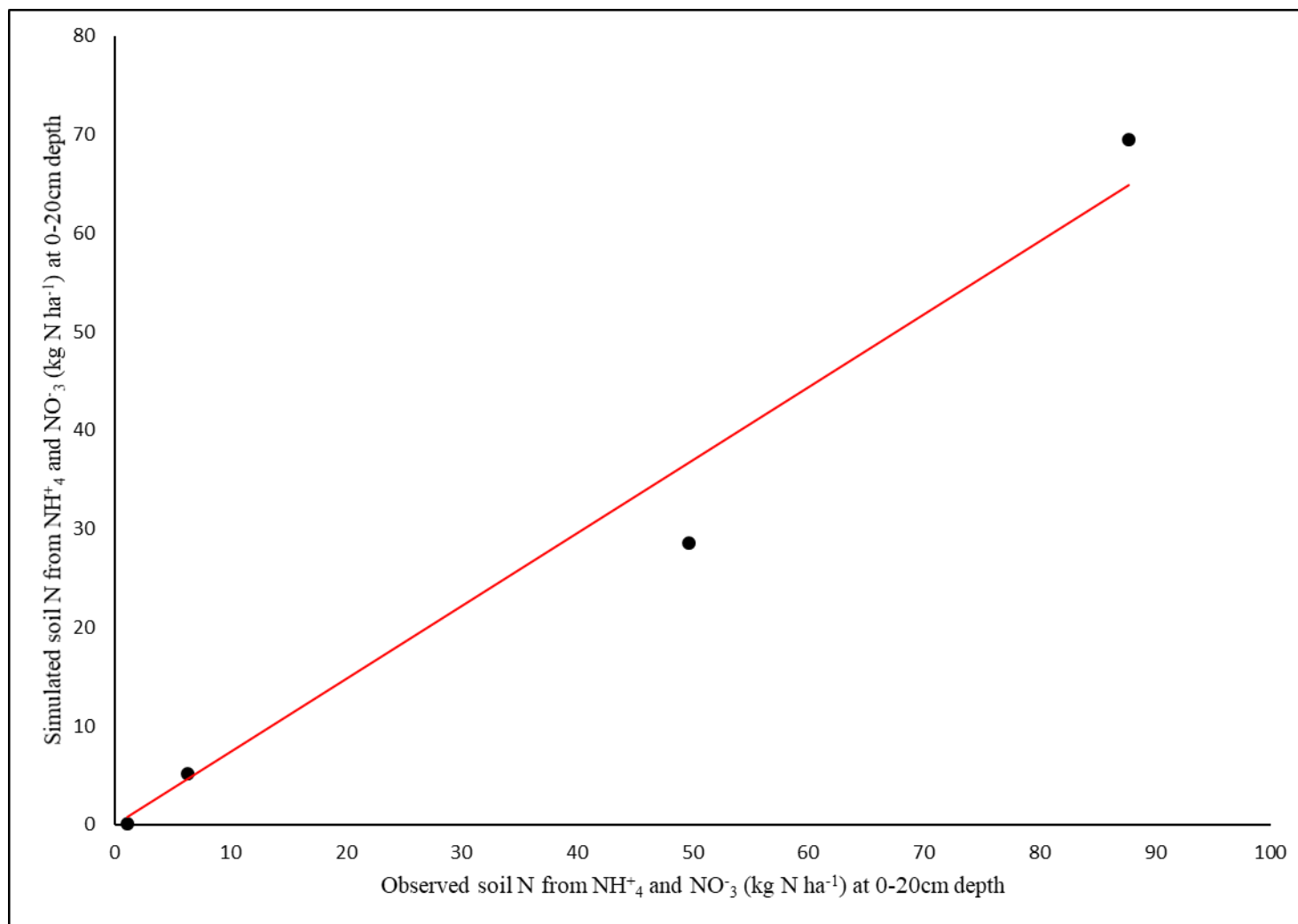
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